1 Quantification of carbon nanotubes in

² environmental matrices: Current capabilities,

³ case studies, and future prospects

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26 ABSTRACT

27 Carbon nanotubes (CNTs) have numerous exciting potential applications and 28 some that have reached commercialization. As such, quantitative measurements 29 of CNTs in key environmental matrices (water, soil, sediment, and biological 30 tissues) are needed to address concerns about their potential environmental and 31 human health risks and to inform application development. However, standard 32 methods for CNT quantification are not yet available. We systematically and 33 critically review each component of the current methods for CNT quantification 34 including CNT extraction approaches, potential biases, limits of detection, and 35 potential for standardization. This review reveals that many of the techniques 36 with the lowest detection limits require uncommon equipment or expertise, and 37 thus, they are not frequently accessible. Additionally, changes to the CNTs (e.g., agglomeration) after environmental release and matrix effects can cause biases 38 39 for many of the techniques, and biasing factors vary amongst the techniques. Five 40 case studies are provided to illustrate how to use this information to inform 41 responses to real-world scenarios such as monitoring potential CNT discharge 42 into a river or ecotoxicity testing by a testing laboratory. Overall, substantial 43 progress has been made in improving CNT quantification during the past ten 44 years, but additional work is needed for standardization, development of 45 extraction techniques from complex matrices, and multi-method comparisons of 46 standard samples to reveal the comparability of techniques.

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48 Introduction

The steady increase in potential applications¹ and production^{1,2} of carbon nanotubes 49 50 (CNTs) and their inevitable release during the life cycle of products has raised questions regarding their potential impact on humans and the environment.^{3,4} CNTs can be 51 52 conceptually understood as rolled up graphitic sheets of hexagonally arranged carbon atoms with sp^2 hybridization. These materials have exceptional mechanical strength as 53 54 well as thermal and electrical conductivity properties that make them ideal for a myriad 55 of potential applications (e.g. construction, environmental, optical, electronic, and biomedical).⁵⁻⁸ The annual production capacity of CNTs reached 2×10^6 kg (2.25 ktons) 56 yr^{-1} in 2011 with an estimated production capacity of 5 x 10⁶ kg (4.5 ktons) yr^{-1} ; this 57 58 change was a 10-fold increase since 2006.¹ With increasing production volume, it is 59 important to determine the potential for biological exposures to CNT during the 60 production, usage, and disposal of CNT-enabled products. The necessary linchpin to 61 quantifying potential CNT exposure, and any risks from it, is the availability of robust 62 analytical methods for quantifying CNTs in complex environmental matrices.⁹ These 63 methods are critical for the assessment of potential CNT exposure, toxicity testing on 64 the potential risks that may occur after exposure, and determination of the environmental fate of CNTs.¹⁰ 65

Analytical techniques to quantify CNTs usually rely on unique physicochemical properties of CNTs that differentiate them from other compounds in relevant media. These approaches leverage the structural, thermal, and electrical properties of CNTs and include spectroscopic,^{11,12,13,14,15} optical,^{16,17} and thermal^{16,14,18} techniques used individually or in combination.^{9,15} Importantly, techniques used for analysis of traditional organic and inorganic toxic chemicals are often not applicable for the following reasons: a) unlike most organic pollutants, CNTs have a distribution of 73 lengths and diameters rather than a single molecular structure and, therefore, mass 74 spectrometry methods, a key tool in current organic analytical methods, generally 75 cannot be used; the large molecular weight of CNTs could potentially challenge mass 76 spectrometric methods too; b) most techniques cannot distinguish between CNTs and 77 naturally occurring black carbon allotropes (e.g., soot or charcoal), which are present at much higher concentrations in the environment than those modeled for CNTs¹⁹; c) 78 79 several other carbon forms are often present in samples (e.g., natural organic matter; 80 NOM) which may interfere with CNT quantification in the sample matrix; and d) the 81 wide range of shapes, sizes, diameters, functional groups, and agglomeration states 82 make it difficult to develop a universal analytical method for quantifying all types of 83 CNTs. In addition, commercially manufactured CNTs may also contain substantial 84 concentrations of metal catalysts, amorphous carbon, and graphitic (non-CNT) 85 nanoparticles (NPs) which may cause biases with some analytical techniques, but are 86 essential for other techniques.²⁰⁻²²

87 While there have been numerous analytical techniques used to quantify CNTs in various matrices,^{4,14-16,23-38} for each technique there have only been a limited number 88 89 of studies, often made by a single laboratory, and thus the robustness of the methods is 90 unknown. In particular, relevant experimental parameters including comprehensive 91 characterization of the CNTs and quantities used for testing and calibration procedures 92 are not always reported. Moreover, failed attempts to apply new methods and 93 techniques or to replicate approaches described in previous studies are often not 94 published, and thus, the limitations of each technique such as potential biases for 95 various matrices (e.g., water or soil with natural (NOM) or soil organic matter (SOM)) 96 are often unclear. Overall, while some recent review papers have focused in part on CNT quantification,^{4,39,40} many critical topics (e.g., interferences in key matrices 97

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98 (environmental, biological, synthetic polymers), and the potential biases with CNT
99 quantification from changes to the CNTs (e.g., oxidation)) related to the development
100 of robust, precise, and reproducible CNT quantification methods have not yet been
101 critically evaluated.

102 This manuscript reviews CNT quantification techniques and evaluates their 103 applicability for different key matrices (water, soil/sediment, tissue) and different types 104 of CNTs (i.e., single-wall carbon nanotubes (SWCNTs) or multiwall carbon nanotubes 105 (MWCNTs)). We report a critical evaluation and comparison among the advantages 106 and limitations of each technique including biases for relevant matrices, biases from 107 physicochemical changes to CNTs in those matrices (i.e., oxidation/degradation, 108 wrapping with organic molecules, and agglomeration), detection limits in various 109 matrices, the potential for standardization, and the types of CNTs that can be analyzed. 110 In addition, methods for extraction or separation of CNTs from different matrices, 111 which may be necessary for sample preparation for some techniques, are enumerated. 112 These quantification, separation, and extraction techniques may also be relevant for 113 quantifying CNT loading in consumer products but the focus of this paper will be on 114 scenarios relevant for assessing the potential environmental risks and fate of CNTs. For 115 example, potential quantification techniques for representative scenarios related to 116 environmental release and potential ecotoxicological effects are discussed. Future 117 research topics to elucidate and improve the analytical performance of these techniques 118 and CNT quantification in general are also highlighted. This paper is intended to serve 119 as a reference to guide scientists in the area of CNT quantification through the selection 120 of an appropriate technique given a type of CNT, sample matrix, and CNT 121 concentration. Given the substantial literature on physicochemical properties and characterization of CNTs,^{41,42} basic background information on these subjects is not 122

provided. While CNTs are also widely known to cause artifacts in many nanotoxicology
assays such as by adsorbing key reagents,^{26,43-45} this manuscript will focus on biases
related to quantification of CNTs and not biases in the measurements of their potential
toxicological effects.

127 Extraction and Separation Procedures for CNTs

128 Numerous techniques have been investigated to extract or separate CNTs from 129 different matrices to overcome quantification limits in complex biological and 130 environmental media (Table 1). In this manuscript, we define "extraction" as the 131 isolation of analytes from a matrix by their physical transition from one phase into 132 another. In contrast, separation means the isolation of analytes from themselves (e.g. 133 differently sized CNTs), or from a matrix within a given phase (e.g. a mobile phase in 134 chromatography or field flow fractionation). Successful extraction methods usually 135 involve the suspension of CNTs in a specific media in which interfering compounds 136 are less soluble, but the converse approach can also be utilized: removing the matrix 137 while leaving the CNTs. However, most reported separation or extraction methods have 138 only been used by a single research group in one or a small number of studies to partly 139 or fully separate CNTs from an environmental matrix (e.g., asymmetric flow field flow fractionation (AF4), matrix digestion, and sonication with surfactants).^{15,23,46} Other 140 141 techniques have not yet been utilized with environmental and biological matrices (e.g., 142 density gradient centrifugation, gel permeation chromatography, capillary 143 electrophoresis, two-polymer phase extraction), but instead have been successfully 144 applied to simpler matrices (e.g., deionized water) or have been used for CNT purification.⁴⁷⁻⁴⁹ These techniques may be valuable for use with environmental and 145 146 biological matrices and are also listed in Table 1. Conversely, there has been more progress with extraction and analysis of fullerenes, another carbon nanomaterial, from
 complex matrices.⁵⁰⁻⁵⁶

149 Currently, many challenges remain in CNT extraction and separation strategies. 150 First, it is unclear to what extent many of these techniques would be applicable for both 151 MWCNTs and SWCNTs given the different properties of these two classes of CNTs, 152 as most methods have only been applied to one or the other. This thought may be 153 extended beyond the number of walls, to include any change in physicochemical 154 properties (e.g., length, internal or external diameter, number of walls, or functional 155 groups). Nevertheless, we expect that separation and extraction techniques may have to 156 be tailored for a specific physicochemical property. For example, a method that can 157 isolate short CNTs from a matrix could be ineffective when used against a population 158 of long, highly entangled CNTs. Second, separation or extraction methods have not yet 159 been applied to CNTs as utilized in potential consumer applications such as in polymer 160 nanocomposite matrices. Given that CNTs will be released into the environment from 161 consumer products, it is important to quantify the release of CNTs from these products 162 after environmental stresses. It may also be important to quantify the concentration of 163 CNTs in the consumer products, such as CNT-containing nanocomposites, to determine 164 the potential quantity that could be released. Given challenges related to collecting and 165 quantifying CNTs released from polymeric nanocomposites, one approach to estimate 166 the quantity of CNTs released is to use a mass balance approach by quantifying the 167 CNT concentration in a product before and after environmentally relevant degradation 168 processes. For example, established methods are needed to extract CNTs from CNT-169 containing nanocomposites before and after the weathering and degradation processes 170 (e.g., due to UV degradation and abrasion) to enable quantification of CNT concentrations.⁵⁷⁻⁶⁰ This will allow scientists to more fully address the complete life 171

172 cycle of nano-enabled consumer products. Finally, extraction or separation procedures 173 may change the physicochemical properties of the CNTs, potentially impacting the 174 reliability of results from analytical methods. One such example is the matrix digestion approach described by Doudrick et al.,²³ which was suitable for subsequent analysis 175 176 using thermal optical transmittance (TOT), but is potentially unsuitable for 177 spectroscopic quantification by Raman scattering, because of concerns that the Raman 178 spectra (e.g., ratio of D to G band) may be altered by the digestion procedure. Overall, 179 although encouraging results have been obtained for a limited number of studies, the 180 overall development of extraction and separation methods for CNTs from matrices for 181 quantitative analyses is still a relatively new area of research.

182 **Quantification techniques**

183 A broad range of techniques have been developed to quantify or identify CNTs 184 in environmentally and biologically relevant matrices (Table 2). In general, the 185 techniques can be sorted into four groups: those that rely on the unique spectroscopic 186 and thermal characteristics of the CNTs (that enable them to be distinguished from the 187 matrix), those that utilize the presence of metal catalyst impurities (associated with the 188 CNTs from the synthesis process), those that require isotopically enriched or depleted 189 CNTs (e.g., with carbon-14 or carbon-13), and finally, microscopic techniques. There 190 are large differences in the sensitivities and applicability of these techniques. Some 191 thermal processes produce detectable gases (CO, CO_2), while others measure radiative 192 heating of a sample. For example, the microwave method involved irradiating CNT 193 containing samples with microwave radiation, wherein the carbon nanotubes absorb the 194 microwave radiation, and the increase in temperature is proportional to the CNT concentration for a given matrix.^{61,62} When comparing different studies, even those 195

using the same quantification technique, there is substantial diversity in thecharacteristics of the CNTs utilized.

198 It is evident from Figure 1 that, while some instruments used in the CNT 199 quantification techniques are commercially available (e.g., UV/vis/NIR spectroscopy 200 and Raman spectroscopy), most of the techniques require uncommon equipment that 201 need to be partially or wholly custom built (e.g., microwave method, photoacoustic and 202 photothermal imaging) or expertise that is not readily available. The use of uncommon 203 instruments in these techniques also poses challenges for commercial ecotoxicity 204 testing facilities to fulfill guidelines for standard methods related to maintaining a consistent exposure concentration.⁶³ While some analytical instruments that can be 205 206 used to quantify CNTs are widely available (e.g., UV/vis spectrophotometry), some of 207 them have significant potential interferences as will be discussed in detail in subsequent 208 To provide one example, challenges related to the use of UV/vis sections. 209 spectrophotometry have recently been described including absorption coefficients 210 dependent on the CNT structure distribution and dispersion method, as well as 211 decreasing absorption coefficients with CNT agglomeration and uncertainty in determining non-CNT from CNT contributions.^{64,65} The lack of robust and widely 212 213 available analytical methods likely contributes to the exclusive use of nominal 214 concentrations to describe the exposure concentration and the absence of reported 215 changes in CNT concentrations during experiments in many nanoecotoxicology 216 studies.

Microscopic techniques can provide unambiguous identification of the CNTs in a complex matrix (e.g., transmission electron microscopy (TEM) analysis using electron energy loss spectroscopy or high resolution TEM),^{27,66} but low or uneven distributions of CNTs on microscopy samples hamper the conversion of the number of

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221 CNTs detected on (several) images to the number/mass concentration of CNTs in a 222 sample. These limitations can be overcome, for matrices without substantial 223 interferences, by using a centrifugation-based method to capture the CNTs from a 224 known volume onto a microscopy sample holder (e.g. TEM grid). Under these 225 conditions, frequency data (number of CNTs per area) can be converted into particle number and mass concentration metrics.⁶⁷⁻⁶⁹ However, when one considers projected 226 227 environmentally relevant concentrations of CNTs (typically ng to µg kg⁻¹ solids),⁷⁰ the 228 likelihood that one captures a CNT onto a microscopy grid with µg-sized environmental 229 samples is exceedingly small. Overall, due to limitations related to the sample 230 preparation issues (low CNT concentration especially compared to other solids, 231 overlapping particles, and uneven distribution of CNTs onto the sample holders), 232 results from electron microscopic techniques remain mainly on a qualitative level, and 233 are currently of limited utility for quantitation.

234 While electron microscopic techniques are very helpful to confirm the identity 235 of CNTs in a matrix if the CNT loading is sufficiently high, reliable controls of the 236 sample matrix without CNTs, the CNTs alone, the sample holder, and any other 237 interferences are needed to avoid false positive or false negative results, but these 238 controls are rarely available for environmental samples. In addition, the amount of time 239 required for sample preparation depends on the samples matrix and greatly varies 240 among techniques. For example, obtaining TEM images suitable for automated image 241 analyses may require that individual CNTs are evenly distributed on a TEM grid and 242 do not overlap with other particles. This often requires elaborate and tailored extraction, 243 dispersion and deposition techniques that are very time intensive to develop. In contrast, 244 sample preparation for hyperspectral imaging microscopy is usually very fast, as liquid 245 samples can be directly cast onto a microscopy slide and subsequently imaged.

However, the current commercial setup lacks the possibility for automated image
acquisition as well as suitable measures to determine the deposited sample volume,
which hampers its quantitative capabilities.

249 Due to the similarities between CNT structure and that of atmospheric soot or 250 carbon black, many analytical techniques that have been used for their extraction or 251 isolation from air, soil, or sediment have been also used to quantify CNTs (e.g., thermal 252 optical transmittance (TOT), chemothermal oxidation at 375 °C (CTO-375), 253 thermogravimetric analysis (TGA), and total organic carbon (TOC)).^{14,16,18,71} While 254 TOT can measure CNTs, custom temperature ramping programs are required for CNTs 255 that differ from standard National Institute for Occupational Safety and Health (NIOSH) methods used for soot analysis on atmospheric samples.¹⁶ Similar 256 257 modifications may also help improve CNT quantification by other thermal techniques 258 such as CTO-375. Sampling of soot in air requires separation from the air, and usually involves filters, impactors or centrifugal separation. Airborne CNTs would likely also 259 260 be captured by these techniques.⁷²⁻⁷⁶

261 All of the quantification techniques are critically assessed in subsequent 262 sections for the potential impact of matrix interferences or interferences from changes 263 that may occur to the CNT in different test systems or the natural environment. For 264 example, the impact of CNT degradation, as has been shown to occur enzymatically and due to interactions with cells and bacteria,⁷⁷⁻⁸³ and oxidation on the performance of 265 266 different analytical methods are evaluated. In addition, the limits of detection (LODs) 267 for these techniques in different media are compared and used to assess the potentially 268 relevant techniques for five case study scenarios. The potential for these techniques to 269 be standardized, a critical issue for regulatory agencies, is also discussed.

270 Evaluation of potential matrix interferences for quantification procedures

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271 Perhaps the principal reason that quantification of CNTs in environmentally 272 relevant matrices is challenging is because of matrix interferences, namely difficulties 273 associated with detecting carbon in a carbon background, especially at modeled average 274 environmental CNT concentrations.^{70,84-86} The matrix characteristics that are most 275 likely to cause interferences are described in detail in Table S1. Overall, natural waters 276 and cell media (e.g., in studies with fish or human cells) have significantly fewer matrix 277 interferences compared to biological tissues, soil/sediment, and released material from 278 nanocomposites. For most spectroscopic measurements, while molecules and 279 suspended particles in natural waters and cell media can potentially scatter 280 incoming/outgoing light thus potentially biasing measurements, methods that account 281 for these effects are generally available; in contrast, separation from the matrix is often 282 needed prior to CNT quantification in tissues, soil/sediment, and fragments released 283 from nanocomposites. For inorganic elemental analysis, having a constant and 284 relatively low background metallic content of the matrix of the same element as the 285 catalyst(s) of the CNTs is most important for all relevant matrices to achieve a low LOD 286 and accuracy. Additionally, the multi-isotopic capability of the inorganic elemental 287 analysis may enable qualitative and/or quantitative isotopic analysis when the isotopic 288 ratios of the catalyst particles differ from those typically observed in the environmental 289 matrix. For single-particle inductively coupled plasma-mass spectrometry (spICP-MS) 290 analysis of CNTs, the background metallic content in nanoparticulate form in matrices 291 is similarly important with regards to the accuracy of the measurement, while low 292 background metallic content in dissolved form is necessary for achieving a low LOD. 293 However, spICP-MS instruments operating at microsecond dwell times can only 294 perform nanoparticle isotopic analysis for detection of two elements, a capability which 295 nevertheless can be used to distinguish naturally occurring NPs from their engineered counterparts.⁸⁷ While spICP-time of flight (ToF)-MS has recently shown the capacity
for multi-element analysis,^{88,89} the size limit of detection was larger for gold and silver
NPs compared to quadrupole-based instruments.⁸⁹ Given the expected small amounts
of the catalysts associated with individual CNTs and challenges associated with
determining the background cut off level for SWCNT analysis using spICP-MS,²² it is
unclear if spICP-ToF-MS will work well for CNT quantification.

302 Thermal techniques often do not show interferences with natural waters and cell 303 media, although there were technique-specific chemicals in these matrices (e.g., peptone in the media for TGA analysis)⁹⁰ that could impact the results. Two key 304 305 considerations for many of the thermal techniques are whether components in the 306 matrix can change the thermal stability of the CNTs and if there is the potential for 307 overlap in the oxidation temperatures of CNTs and combustible components of the 308 matrix. Thermal techniques could generally work in all matrices but the detection limit 309 will be higher in matrices with more interferences as will be discussed in a subsequent 310 section. Lower LODs may be achievable by first extracting the CNTs or decreasing the 311 bias from other forms of organic carbon.

Quantification of CNTs (and other carbon nanomaterials⁹¹⁻⁹⁵) via isotopic 312 313 labelling generally has fewer interferences than the other techniques, but obtaining 314 isotopically enriched CNTs is typically challenging and/or expensive. Furthermore, this 315 approach is only relevant for laboratory studies, not for detecting CNTs released into 316 the environment. A related strategy, labeling CNTs with coatings containing a radioisotope, was used in many early biodistribution studies in the biomedical field,⁹⁶⁻ 317 ⁹⁸ but has not been used in environmental or ecotoxicological studies. The challenge 318 319 with this approach is that the accuracy of any measurement is contingent upon the 320 radioactive tracer remaining associated with the CNTs.

Natural abundance, stable isotopic measurements (e.g., carbon-13)²¹ face 321 322 similar limitations in that they require a CNT-free sample to which one can compare 323 the isotopic composition in order to deploy the technique quantitatively. In laboratory 324 studies, this is possible and more economically viable than radiolabeling techniques, 325 but one has to carefully select CNT-free controls for quantifying CNTs in 326 environmental samples. Furthermore, while the initial label is more expensive, the 327 analytical techniques required to trace a carbon-14 label (i.e., liquid scintillation 328 counting) are facile compared to the expert preparatory and analytical equipment 329 required to trace natural-abundance isotopes (i.e., much lower levels of either carbon-330 14 or carbon-13 require accelerator mass spectrometers and isotope ratio mass 331 spectrometers, respectively, and each with closed-tube-combustion preparation 332 upstream). Nevertheless, the carbon source for SWCNTs produced using the high 333 pressure carbon monoxide (HiPco) process is usually biomethane,²¹ which has a strong 334 naturally depleted carbon-13 signature, and such CNTs would be good candidates for using natural abundance, stable isotopic measurements.⁷⁸ 335

336 Evaluation of potential bias from changes to the CNTs

337 In addition to interferences from different environmentally and biologically 338 relevant matrices, changes that may occur to CNTs while in these matrices can also 339 cause interferences for many of the quantification techniques (Table S2). The extent to 340 which agglomeration, degradation, and wrapping by other molecules occurs depends 341 on the physicochemical properties of the CNTs and of the matrix. It is well known that 342 CNTs will agglomerate in waters with sufficient ionic strength if they are not stabilized 343 through, for example, a surfactant and that CNTs have a large capacity to adsorb natural organic matter.⁹⁹⁻¹⁰² With regards to CNT agglomeration, while most techniques are 344 345 sensitive to this change (e.g., most thermal techniques, Raman, NIRF, UV/vis/NIR

346 absorbance, and spICP-MS), some are not impacted by it (e.g., inorganic element 347 analysis) or may even be enhanced (e.g., hyperspectral imaging). Potential interference 348 from CNT agglomeration may result in, for example: a) changes to the intensity or peak 349 wavelengths in the spectrophotometry signals; b) shifts in the thermal stability of the 350 CNTs, which could prevent separation from other components in the matrix, such as 351 black carbon soot; or c) hindering uniform distribution on a filter prior to analysis by 352 TOT. Agglomeration may also increase the heterogeneity and affect representativeness 353 of the subsamples in a matrix, which could lead to increased uncertainty. However, 354 larger subsamples could help lower the uncertainty when feasible.

355 The literature shows variable results on the degradation of CNTs in 356 environmental matrices. In some studies, degradation of carbon-14 labeled CNTs by enzymes or bacteria has been shown to be slow or not detectable ^{77,78,103} except under 357 specific situations with a special microbial consortium.⁷⁷ In contrast, studies assessing 358 the degradation of non-carbon-14 labeled CNTs have often shown substantial 359 360 degradation.⁸² The cause of this discrepancy is unclear. Studies on the photodegradation 361 of CNTs have shown significant modifications to their surface structure or the loss of 362 fluorescence under some experimental conditions.^{104,105} Thus, it is reasonable to 363 assume that some degree of degradation could occur with CNTs in surface waters if 364 they stay suspended for a sufficiently long period. Almost all quantification techniques 365 are sensitive to CNT degradation and oxidation, although the degree of oxidation 366 needed before it impacts quantification varies among techniques. One exception is 367 carbon-14 analysis, which is not impacted by oxidation. In contrast, the degree of oxidation can directly impact CNT thermal properties and potentially the capacity to 368 369 differentiate between CNTs and other forms of carbon present in the matrix using many 370 of the thermal based techniques.

371 Wrapping of organic molecules around CNTs, such as proteins or NOM, may 372 also impact most quantification techniques. Many of the potential changes that could 373 cause biases, such as decreased signal intensity of a spectroscopic measurement or a 374 change in the thermal stability of CNTs for thermal measurements, are similar to those 375 discussed for degradation. However, the reason behind these changes is from the impact 376 of the coating on the CNT properties rather than a change to the core CNT material 377 itself as would occur during degradation. One challenge in discussing the potential bias 378 from organic molecules wrapping around CNTs, and also agglomeration and 379 oxidation/degradation, is that the magnitude of the bias relates partly to the degree of 380 agglomeration, oxidation, and the quantity of organic molecules associated with the 381 CNTs. It is possible to foresee examples when these changes in the environmental 382 matrices could have a bias, but it is challenging to quantify the magnitude of the 383 expected bias without information about the sample system (e.g., aqueous phase NOM 384 concentrations can range between 5 mg/L and 50 mg/L) or the extent of oxidation. This 385 information about the sample system or magnitude of likely changes could allow one 386 to account for biases.

387 Being aware of the potential biases present in a sample from these changes to 388 the CNTs and/or carrier matrix will support researchers in determining to what extent 389 these factors may impact their measurements. However, it might be challenging to get 390 this kind of information from samples with low CNT concentrations when there is a 391 low signal to noise ratio. Environmentally-relevant information on the rate of CNT modifications (e.g., oxidation) by environmental processes is limited, 77,103,106-108 and 392 393 systematic studies of those processes would be an enormous benefit to parallel efforts 394 to quantify CNTs in the environment. While leaching of metal catalysts from the CNTs 395 in environmental matrices is not explicitly covered in the above changes to the CNTs,

it could dramatically impact analyses using spICP-MS or elemental analysis. The potential for changes in the catalyst particles associated with the CNTs in environmental matrices is the primary reason that these techniques are not more broadly used despite their low LODs.

400 *Detection limits of quantification techniques*

401 The LOD for CNT quantification is one of the most critical performance metrics 402 required to compare the various techniques. However, the definition of the LOD 403 depends partly on how the CNT mass in a given sample is determined. The most 404 common approach is for the whole sample, including CNTs, catalyst particles, and any 405 carbonaceous impurities, to be included in the CNT mass used. It is possible instead to 406 only use the CNTs themselves, at least for SWCNTs where, after purification procedures, the properties are more clearly distinguishable and high quality separation 407 408 techniques exist.¹⁰⁹ While additional metrics such as number or surface area concentrations are highly desired,^{63,110} the LOD values provided here are for mass 409 410 concentrations.

411 There are two different approaches for determining the necessary LOD for 412 quantifying contaminants in the environment. The first requires that the LOD is 413 adequate for quantification of the contaminant at concentrations that may have harmful 414 effects. An alternative requirement is for the analytical techniques to quantify the 415 contaminant at the concentration that it is determined or estimated to be present in the 416 environment. We have compared the LODs for the various analytical techniques using 417 both approaches through comparing the LODs to a species sensitivity distribution for 418 CNT acute toxicity to pelagic organisms (Figure 2) and to modeled environmental 419 concentrations (Figure 3). Several trends are evident from reviewing these figures. 420 First, the LODs in water span several orders of magnitude with some techniques only

421 capable of quantifying CNTs in samples with concentrations greater than 10 mg/L (e.g., 422 gravimetric measurements), while the most sensitive techniques can detect 423 concentrations between 0.1 μ g/L and 1 μ g/L (e.g., spICP-MS) (Figure 3). Second, the 424 lowest LOD values are for pristine water samples and increase with higher amounts of 425 potential interferences in the matrix. Higher LODs are observed when NOM is present 426 in waters, and even higher LODs are typically achieved when using CNT quantification 427 techniques in soils, sediments, and biological tissues. Third, multiple techniques appear 428 capable of quantifying CNTs at concentrations relevant for stock suspensions (e.g., 10 mg L^{-1} to 100 mg L^{-1}) that could be used for pelagic aquatic toxicity testing (Figure 2). 429 430 As discussed in more depth in a case study, some techniques could also be used to 431 quantify the initial exposure concentration for ecotoxicity testing and the concentration 432 after the experiment concludes. Fourth, the LODs are often orders of magnitude higher 433 than the average modeled environmental concentration, but some are within the range 434 of modeled sediment concentrations despite the lower LODs for CNT quantification in 435 sediments. This suggests that it may be feasible to quantify CNTs in the environment 436 under certain conditions. Overall, these figures can be used to assess which methods 437 may offer suitable techniques for an intended purpose, as is described in more detail in 438 the case studies. Alternatively, extraction or separation techniques (see above) may be 439 necessary to selectively isolate and concentrate the CNTs prior to analysis.

440 Potential for standardization

There are numerous reference materials (RM; e.g., UV/vis spectroscopy calibration standards) and standard methods that can support the standardization of CNT quantification techniques (Table S3). In addition, there are multiple CNT RMs and representative test materials (Table S4); RMs have assigned values for certain properties, whereas representative test materials are only guaranteed to be stable and 446 homogeneous with respect to one or more specified properties but may be used in the 447 development of test methods which assess properties other than those for which stability and homogeneity have been shown.¹¹¹ Currently, three RMs are available for 448 SWCNTs, while MWCNTs are only available as representative test materials. The 449 450 careful characterization of the CNT RMs may be useful for the standardization of 451 numerous techniques, given the wide range of properties that have been certified (i.e., 452 the sources of uncertainty are thoroughly understood and the certified values have 453 meaningful metrological traceability) or for which information values are provided 454 (i.e., the sources of uncertainty are not fully understood or a limited number of analyses 455 were performed). Standardized methods are also already available for characterization 456 of CNTs (e.g., Raman spectroscopy and NIR fluorescence characterization) which could be modified to develop standard methods for CNT quantitation.^{42,112-118} In 457 458 addition, a modified version of a NIOSH standard method for use of TOT for elemental 459 carbon analysis (NIOSH Method 5040) could potentially be used for CNT 460 quantification. However, the robustness of this method for CNTs will still need to be 461 evaluated for different matrices. Extraction and separation procedures also need to be 462 standardized but are not addressed in this section due to the limited number of studies 463 on this topic. Research topics that would support the standardization of these techniques 464 are described in the Future Research Topics section.

465 Case studies

In this section, five case studies will be used to illustrate how the quantitative methods described in this manuscript could be utilized to address hypothetical situations requiring CNT quantitation. The scenario for the first two case studies is that scientists are asked to determine whether the concentration of CNTs in a stream receiving effluent from a treatment plant where CNTs may be released is above 500 μg

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 L^{-1} ; this concentration was chosen because it is approximately 50 % of the lowest LC₅₀ 471 472 value of the species sensitivity distribution shown in Figure 2. This scenario will be 473 discussed in the context of whether the CNT characteristics (e.g., SWCNT or MWCNT, 474 catalyst materials, and thermal properties) are known *a priori* or not. In the third case 475 study, scientists will be trying to measure the exposure concentration to organisms 476 during a laboratory ecotoxicity experiment in a water only system with an organism 477 that has an EC_{50} value (the concentration at which 50 percent of the organisms are affected) of 10 mg L⁻¹ and the lowest concentration tested is 1 mg L⁻¹. In the fourth case 478 479 study, CNTs with known characteristics are accidentally released into a lake, and 480 scientists are asked to determine the concentration in the lake sediment. In the fifth case 481 study, "OECD Test 305: Bioaccumulation in fish: aqueous and dietary exposure" is 482 performed using a known type of CNTs and the scientists need to quantify the 483 concentration in the fish tissues.

484 *Case I: CNTs with known characteristics are released into a river*

485 First, identify the techniques that may have LODs better than 500 μ g L⁻¹ using 486 Figure 3: UV/vis spectroscopy, inorganic elemental analysis, spICP-MS, NIRF, Raman 487 spectroscopy, TOT, and carbon-14 labeling. Electron microscopy should, in principle, 488 be able to detect CNTs at these concentrations, but it may be challenging to identify 489 CNTs amidst the other particulate matter, and quantification will be challenging as 490 discussed above. Of particle risk is the ability to collect a representative sample where 491 the TEM thin section actually contains a statistically significant number of CNTs. 492 Nevertheless, electron microscopy could be used for a qualitative assessment or to 493 confirm the presence/absence of CNTs based on results from the quantitative analysis. 494 Among the quantitative techniques, the choice of which technique to employ first would 495 depend on numerous factors such as their availability and if the unique properties of the

496 CNTs of interest may eliminate some of the analytical techniques from consideration 497 (e.g., quality assurance (QA), techniques only applicable for SWCNTs would not be 498 relevant for MWCNT quantification). For example, carbon-14 labeling would not be 499 relevant for field measurements, while NIRF would only be applicable for SWCNTs.¹⁵ 500 In addition, Raman spectroscopy analysis would require preconcentration of the sample to yield the desired LOD which may be challenging.¹³ Next, the properties of the river 501 502 water prior to the discharge location (e.g., thermal profile, elemental composition and 503 organic matter concentration of the water) could be evaluated to assess what biases may 504 be encountered during CNT quantification for various techniques. If it is possible to 505 obtain the CNTs of interest, a next step would be to prepare a CNT dispersion, mix the 506 dispersion with stream water prior to the location of discharge, and then analyze the 507 water using the quantification technique(s) to determine relevant QA/quality control 508 (QC) characteristics such as the LOD, reproducibility, bias, signal to noise ratio, and 509 linearity of calibration curve. It may also be important to test the stability of the CNT 510 in the water prior to the discharge location to assess if agglomeration or oxidation of 511 the CNT could cause a bias; if agglomeration causes a significant bias, it may be 512 possible to disperse the samples such as by adding a surfactant or sonicating the sample. 513 If the QA/QC characteristics are sufficient to provide the needed level of statistical 514 significance for the quantification measurement, the final step would be to analyze the 515 test samples.

516 Case II: CNTs with unknown characteristics are released into a river

517 The process is substantially more complicated if characteristics of the CNT to 518 be detected are unknown. First, it would be helpful to obtain water samples before and 519 after the point source discharge location. It would then be possible to do some 520 measurements to try to determine if characteristics of the river water reflective of CNT 521 characteristics are changed. For example, an elemental analysis or spICP-MS analysis 522 of the river waters could be conducted to assess if uncommon elements (e.g., yttrium) 523 or ratios of elements (e.g., cobalt to molybdenum) often used for CNT catalysts are 524 present at different concentrations before and after the location of discharge; measuring 525 these samples before and after filtering could reveal if the metals are associated with 526 particles such as CNTs. One distinct advantage of the metal analysis techniques is that 527 the LODs for many of these elements are orders of magnitude better than the limit of 528 detection needed for the CNTs (Figure 3). This information supported by other 529 characterization techniques (e.g., TEM analysis to assess if SWCNTs or MWCNTs can 530 be identified) could help determine the type of CNT being used. An alternate first step 531 would be to obtain a sample directly at the discharge location and conduct these 532 analyses. The advantage of this approach is that there would not be dilution of the 533 CNTs, but the matrix may be substantially more complex (e.g., wastewater treatment 534 plant effluent). A next step is to spike known concentrations of the specific CNT if 535 identified, or alternatively RM SWCNTs and representative test material MWCNTs, 536 into the river water prior to the discharge location and determine the QA/QC 537 characteristics for the selected techniques and the extent to which agglomeration or 538 oxidation could influence the results. If acceptable results can be obtained with the 539 specific CNT (if identified) or the RM CNTs, then analysis can be conducted on the 540 river sample after the location of discharge.

541 *Case III: Laboratory Ecotoxicity Study*

542 The third case study involves a laboratory ecotoxicity experiment during which 543 the concentration remaining suspended during the experiment needs to be quantified. 544 Depending upon what organism is tested, there may be interferences such as algae or 545 bacteria which remain suspended and have CNTs associated with them. If it is

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546 straightforward to separate the test organisms from the media with suspended CNTs, 547 numerous techniques may be applicable for quantifying the initial CNT concentration 548 in suspension ($\geq 1 \text{ mg L}^{-1}$) (see Figure 2). The techniques available to determine the 549 change in concentration during the experiment depend on the LOD needed for these 550 measurements. For example, if it is unlikely that the CNT will settle during the 551 experiment, numerous techniques would enable measurements to show that the 552 concentration remained within 20 % of the initial concentration, the desired maximum 553 concentration loss indicated in many OECD tests.⁶³ However, if substantial settling 554 occurs, it is necessary to determine the lowest detection limit needed (e.g., 0.1 mg L⁻¹ 555 to quantify a loss in concentration of 90 % of the initial concentration). When measuring 556 the CNT concentration dispersed in tests with suspended unicellular organisms or small 557 multicellular organisms (e.g., Tetrahymena thermophila), the cells themselves may 558 cause biases or require the extraction of the CNTs. It is also unclear if CNTs that are 559 suspended but associated with cells should be counted as part of the total suspended 560 concentration. Nevertheless, many techniques could likely still be used to quantify the 561 total suspended concentration but control experiments to test for potential biases from 562 the cells and the matrix would need to be conducted prior to starting the experiment.

563 Case study IV: Quantification of CNTs with known characteristics in lake sediment

Quantifying CNTs in sediments is substantially more difficult than in water samples. As shown in Figure 3, the LODs for most techniques are at least an order of magnitude higher in soils and sediments compared to in waters. To quantify CNTs in sediments, a first step would be to obtain "clean" sediment from another water body ideally with similar sediment characteristics. Because the CNT type is known in this case study, it is possible to spike this clean sediment with CNTs and then assess the quality of the analytical results (*e.g.*, linearity, LOD, etc.). The suitable techniques for 571 this analysis will depend upon instrument availability, the type of CNT (e.g., NIRF after 572 CNT extraction has been shown to be a valuable technique for analysis of SWCNTs in sediments¹⁵ but is not applicable to MWCNTs), and the estimated range of probable 573 574 CNT concentrations in the sediment. If satisfactory LODs are not available for the 575 available techniques in the reference sediment, it may be necessary to investigate extraction or separation methods to decrease the LOD (e.g., ^{15,46}). Given the low 576 detection limits obtained using NIRF after extraction (62 µg/kg),¹⁵ challenges with 577 578 obtaining a better LOD are likely only to be problematic for MWCNTs unless the 579 SWCNTs are oxidized or modified to the extent that NIRF is not applicable or NIRF is 580 not available for sample analysis.

581 Case study V: Quantification of CNT in fish after a standard toxicity test

582 Assessing potential bioaccumulation of chemicals in organisms is an important 583 component of risk assessment of chemicals. One frequently used test is OECD method 305: Bioaccumulation in fish: aqueous and dietary exposure.¹¹⁹ Again, the LODs for 584 585 quantifying CNTs in organism tissues are greater than those in water, yet similar to the 586 LODs for soils and sediments (Figure 3). While the whole fish is usually analyzed in 587 this method, it may be beneficial to test the CNT biodistribution in addition to the total 588 concentration in the fish. This is important because CNT translocation across the gut 589 tract is rarely observed in ecotoxicological studies.^{27,31,120,121} If the biodistribution of 590 SWCNTs is evaluated, then the technique with the best LOD is NIRF microscopy which has been reported to detect individual SWCNTs.^{31,121} If this instrument is not 591 592 available, Raman microscopy and electron microscopy can be used to assess 593 biodistribution of CNTs in organisms although it is important to carefully avoid artifacts;^{27,43,122,123} however, one should note that G/D ratios are strongly influenced by 594 any sp^2 or sp^3 hybridized carbons present in the organism for Raman microscopy 595

596 analysis. Other microscopic approaches such as photothermal/photoacoustic imaging 597 have also been successfully used to assess the distribution of CNTs in plants, yet are infrequently available (Figure 1).²⁴ To quantify the total concentration of CNTs in the 598 fish, it is possible to use NIRF microscopy for SWCNTs,³¹ but extraction from the fish 599 600 tissue will likely be needed for MWCNTs. An extraction procedure has been published for MWCNTs in rat lungs followed by quantification using TOT,²³ but this approach 601 602 has not yet been used in tandem with other quantification techniques or with fish tissues. 603 If carbon-14 labeled CNTs are available, assessing uptake by and biodistribution in fish through carbon-14 labeling is a viable approach.¹²⁴ The microwave method has also 604 605 shown promise for detecting MWCNTs in biological samples (e.g., earthworms) but requires custom built equipment.^{62,125} 606

607 Future Research Topics

608 The analysis that we present here on the current state of the science with regards 609 to quantification of CNTs in matrices relevant for nanotechnology environmental health 610 and safety measurements also reveals several key future research topics to move this 611 field forward. First, most of the quantification techniques developed for aqueous 612 environments will have potential biases or a higher LOD in complex matrices such as 613 soils and biological tissues. Thus, the continued development of CNT extraction and 614 separation procedures for environmental and biological matrices is a critical topic for 615 additional research. Nevertheless, addressing the quality of the CNT separation depends 616 in part on the robustness and precision of the subsequent analytical techniques, which 617 also need to be improved. Second, sensitivity analyses of techniques can provide 618 relevant information regarding the robustness of an experimental procedure to minor 619 changes to a protocol and the contributions of various steps to the total uncertainty of 620 the result. This approach and related approaches such as cause-and-effect analysis can 621 highlight which steps of a protocol need to be carefully followed to ensure a reliable result and which steps are less critical.¹²⁶ Third, interlaboratory comparisons, where 622 623 multiple laboratories use the same protocol, are needed to standardize the more mature 624 techniques and extraction and separation procedures. While it is necessary to assess 625 many topics related to analytical precision of a single laboratory (e.g., within and 626 between operator variability, instrument to instrument variability, day-to-day 627 variability, all contributing to the within-laboratory repeatability), interlaboratory 628 comparisons can provide unique information about the comparability of results among 629 laboratories (i.e., between-laboratory reproducibility) and potential factors in the 630 protocols that need to be controlled to standardize the procedure. Such information is 631 needed to provide estimates of the bias and precision of an analytical method. Fourth, 632 analyzing an individual or set of homogenized test samples using multiple techniques 633 will be helpful in highlighting method specific biases and the comparability of results among methods (e.g., similarly to a black carbon quantification ring trial¹²⁷). This 634 635 differs from interlaboratory comparisons in that a single sample is analyzed by multiple 636 techniques, as opposed to different laboratories using the same technique and test 637 method. Similar results among orthogonal techniques would lead to greater confidence 638 in the results of the methods while different results could yield insights into biases, 639 strengths, and limitations of different methods. For example, in a recent study on the 640 fate of SWCNTs in a mesocosm, an experimental setup designed to simulate the natural 641 environment that often includes multiple species and which has been used in several nanotoxicity studies,^{128,129} both NIRF and elemental analysis were used on the same 642 samples.²⁹ The agreement among these methods suggested that elemental analysis may 643 644 be a useful approach in these complex matrices if the catalysts used to synthesize the CNTs are of an element with low concentrations in the matrix (e.g., Mo).²⁹ A similar 645

646 approach could be used to compare among different extraction or separation techniques with a single sample. Fifth, isotopically enriched or depleted $CNTs^{21,78}$ could be used 647 648 to help develop other orthogonal techniques given that isotopic techniques often have 649 the fewest biases for many of the matrices and changes that could occur to the CNTs in 650 these matrices. Such an approach was used by Schierz et al. to develop the NIRF 651 technique for quantification of SWCNTs in sediments after extraction by also testing the extraction procedure with carbon-14 labeled SWCNTs.¹⁵ Sixth, using extraction 652 653 and/or separation techniques in combination such as AF4 followed by capillary 654 electrophoresis could be another promising avenue for future research. Lastly, almost 655 all quantitative techniques require known CNTs to yield information about their 656 characteristic information (e.g., thermal profile, metal catalyst, impurities, NIR spectra, 657 and Raman signature). Additional work is needed to develop techniques for 658 quantification of unknown CNTs in an environmental or biological matrix. Along these 659 lines, the impact of CNT heterogeneities (e.g., different lengths) on their quantification 660 could also be helpful.

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672 **Supporting Information**

- 673 Tables describing potential matrix interferences and interferences from changes to the
- 674 CNTs on selected CNT quantification techniques, standard reference material carbon
- 675 nanotubes, and standards and references materials related to standardization of carbon
- 676 nanotube quantification techniques. This information is available free of charge via the
- 677 Internet at http://pubs.acs.org.

678 References

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680 1. De Volder, M. F.; Tawfick, S. H.; Baughman, R. H.; Hart, A. J., Carbon nanotubes: present and future commercial applications. Science 2013, 339, 681 682 (6119), 535-9. 683 2. Piccinno, F.; Gottschalk, F.; Seeger, S.; Nowack, B., Industrial production 684 quantities and uses of ten engineered nanomaterials in Europe and the world. J. 685 *Nano. Res.* **2012**, *14*, (9), Article number 1109. Helland, A.; Wick, P.; Koehler, A.; Schmid, K.; Som, C., Reviewing the 686 3. environmental and human health knowledge base of carbon nanotubes. Environ. 687 688 Health Perspect. 2007, 115, (8), 1125-1131. 689 Petersen, E. J.; Zhang, L. W.; Mattison, N. T.; O'Carroll, D. M.; Whelton, A. J.; 4. 690 Uddin, N.; Nguyen, T.; Huang, Q. G.; Henry, T. B.; Holbrook, R. D.; Chen, K. L., 691 Potential release pathways, environmental fate, and ecological risks of carbon 692 nanotubes. Environ. Sci. Technol. 2011, 45, (23), 9837-9856. 693 5. Mauter, M. S.; Elimelech, M., Environmental applications of carbon-based 694 nanomaterials. Environ. Sci. Technol. 2008, 42, (16), 5843-5859. 695 Shen, M. W.; Wang, S. H.; Shi, X. Y.; Chen, X. S.; Huang, Q. G.; Petersen, E. J.; 6. 696 Pinto, R. A.; Baker, J. R.; Weber, W. J., Jr., Polyethyleneimine-mediated 697 functionalization of multiwalled carbon nanotubes: Synthesis, characterization, 698 and in vitro toxicity assay. J. Phys. Chem. C 2009, 113, (8), 3150-3156. 699 Shi, X. Y.; Wang, S. H.; Shen, M. W.; Antwerp, M. E.; Chen, X. S.; Li, C.; 7. 700 Petersen, E. J.; Huang, O. G.; Weber, W. J., Jr.; Baker, J. R., Multifunctional 701 dendrimer-modified multiwalled carbon nanotubes: Synthesis, characterization, 702 and in vitro cancer cell targeting and imaging. *Biomacromolecules* **2009**, *10*, (7), 1744-1750. 703 704 8. Lee, J.; Mahendra, S.; Alvarez, P. J. J., Nanomaterials in the construction 705 industry: A review of their applications and environmental health and safety 706 considerations. ACS Nano 2010, 4, (7), 3580-3590. 707 Plata, D. L.; Ferguson, P. L.; Westerhoff, P., Express It in Numbers: Efforts 9. 708 to Quantify Engineered Nanoparticles in Environmental Matrices Advance. 709 Environ. Sci. Technol. 2012, 46, (22), 12243-12245.

710 10. Selck, H.; Handy, R. D.; Fernandes, T. F.; Klaine, S. J.; Petersen, E. J., 711 Nanomaterials in the aquatic environment: A European Union-United States 712 perspective on the status of ecotoxicity testing, research priorities, and 713 challenges ahead. *Environmental Toxicology & Chemistry* **2016**, in press. 714 Bahr, J. L.; Mickelson, E. T.; Bronikowski, M. J.; Smalley, R. E.; Tour, J. M., 11. 715 Dissolution of small diameter single-wall carbon nanotubes in organic solvents? 716 Chem. Commun. 2001, (2), 193-194. 717 Cherukuri, P.; Bachilo, S. M.; Litovsky, S. H.; Weisman, R. B., Near-infrared 12. 718 fluorescence microscopy of single-walled carbon nanotubes in phagocytic cells. *J.* 719 Am. Chem. Soc. 2004, 126, (48), 15638-15639. 720 Lopez-Lorente, A. I.; Simonet, B. M.; Valcarcel, M., Determination of 13. 721 carboxylic SWCNTs in river water by microextraction in ionic liquid and 722 determination by Raman spectroscopy. *Talanta* **2013**, *105*, 75-79. 723 Plata, D. L.; Reddy, C. M.; Gschwend, P. M., Thermogravimetry-Mass 14. 724 Spectrometry for Carbon Nanotube Detection in Complex Mixtures. Environ. Sci. 725 Technol. 2012, 46, (22), 12254-12261. 726 15. Schierz, A.; Parks, A. N.; Washburn, K. M.; Chandler, G. T.; Ferguson, P. L., 727 Characterization and Quantitative Analysis of Single-Walled Carbon Nanotubes 728 in the Aquatic Environment Using Near-Infrared Fluorescence Spectroscopy. 729 Environ. Sci. Technol. 2012, 46, (22), 12262-12271. 730 Doudrick, K.; Herckes, P.; Westerhoff, P., Detection of Carbon Nanotubes 16. in Environmental Matrices Using Programmed Thermal Analysis. Environ. Sci. 731 732 Technol. 2012, 46, (22), 12246-12253. 733 Hyung, H.; Fortner, J. D.; Hughes, J. B.; Kim, J. H., Natural organic matter 17. 734 stabilizes carbon nanotubes in the aqueous phase. Environ. Sci. Technol. 2007, 735 41, (1), 179-184. 736 Sobek, A.; Bucheli, T. D., Testing the resistance of single- and multi-walled 18. 737 carbon nanotubes to chemothermal oxidation used to isolate soots from 738 environmental samples. *Environ. Pollut.* **2009**, *157*, (4), 1065-1071. 739 Koelmans, A. A.; Nowack, B.; Wiesner, M. R., Comparison of manufactured 19. 740 and black carbon nanoparticle concentrations in aquatic sediments. *Environ* 741 Pollution 2009, 157, (4), 1110-6. 742 Mansfield, E.; Kar, A.; Hooker, S. A., Applications of TGA in quality control 20. of SWCNTs. Anal. Bioanal. Chem. 2010, 396, (3), 1071-1077. 743 744 21. Plata, D. L.; Gschwend, P. M.; Reddy, C. M., Industrially synthesized single-745 walled carbon nanotubes: compositional data for users, environmental risk 746 assessments, and source apportionment. Nanotechnol. 2008, 19, (18), Article 747 Number: 185706. 748 Reed, R. B.; Goodwin, D. G.; Marsh, K. L.; Capracotta, S. S.; Higgins, C. P.; 22. 749 Fairbrother, D. H.; Ranville, J. F., Detection of single walled carbon nanotubes by 750 monitoring embedded metals. Environ. Sci. Proc. Imp. 2013, 15, (1), 204-213. 751 23. Doudrick, K.; Corson, N.; Oberdörster, G.; Eder, A. C.; Herckes, P.; Halden, 752 R. U.; Westerhoff, P., Extraction and Quantification of Carbon Nanotubes in 753 Biological Matrices with Application to Rat Lung Tissue. ACS Nano 2013, 7, (10), 754 8849-8856. 755 24. Khodakovskava, M. V.; de Silva, K.; Nedosekin, D. A.; Dervishi, E.; Biris, A. 756 S.; Shashkov, E. V.; Galanzha, E. I.; Zharov, V. P., Complex genetic, photothermal, 757 and photoacoustic analysis of nanoparticle-plant interactions. Proc. Natl. Acad.

758 *Sci. U.S.A.* **2011,** *108,* (3), 1028-1033.

759 Petersen, E. J.; Pinto, R. A.; Zhang, L.; Huang, Q. G.; Landrum, P. F.; Weber, 25. 760 W. J., Effects of polyethyleneimine-mediated functionalization of multi-walled 761 carbon nanotubes on earthworm bioaccumulation and sorption by soils. Environ. 762 Sci. Technol. 2011, 45, (8), 3718–3724. 763 Petersen, E. J.; Pinto, R. A.; Mai, D. J.; Landrum, P. F.; Weber, W. J., Jr., 26. 764 Influence of Polyethyleneimine Graftings of Multi-Walled Carbon Nanotubes on 765 their Accumulation and Elimination by and Toxicity to Daphnia magna. Environ. 766 Sci. Technol. 2011, 45, (3), 1133-1138. 767 27. Edgington, A. J.; Petersen, E. J.; Herzing, A. A.; Podila, R.; Rao, A.; Klaine, S. 768 J., Microscopic investigation of single-wall carbon nanotube uptake by Daphnia 769 magna. *Nanotoxicology* **2014**, *8*, (S1), 2-10. 770 Zhang, L.; Petersen, E. J.; Huang, Q. G., Phase distribution of ¹⁴C-labeled 28. 771 multiwalled carbon nanotubes in aqueous systems containing model solids: Peat. 772 *Environ. Sci. Technol.* **2011**, *45*, (4), 1356-1362. 773 29. Schierz, A.; Espinasse, B.; Wiesner, M. R.; Bisesi, J. H.; Sabo-Attwood, T.; 774 Ferguson, P. L., Fate of single walled carbon nanotubes in wetland ecosystems. 775 *Environmental Science-Nano* **2014**, *1*, (6), 574-583. Ferguson, P. L.; Chandler, G. T.; Templeton, R. C.; Demarco, A.; Scrivens, W. 776 30. 777 A.; Englehart, B. A., Influence of sediment-amendment with single-walled carbon 778 nanotubes and diesel soot on bioaccumulation of hydrophobic organic 779 contaminants by benthic invertebrates. Environ. Sci. Technol. 2008, 42, (10), 780 3879-3885. 781 31. Bisesi, J. H.; Merten, J.; Liu, K.; Parks, A. N.; Afrooz, A.; Glenn, J. B.; Klaine, S. 782 J.; Kane, A. S.; Saleh, N. B.; Ferguson, P. L.; Sabo-Attwood, T., Tracking and 783 Quantification of Single-Walled Carbon Nanotubes in Fish Using Near Infrared 784 Fluorescence. Environ. Sci. Technol. 2014, 48, (3), 1973-1983. 785 Hanna, S. K.; Miller, R. J.; Lenihan, H. S., Deposition of carbon nanotubes by 32. 786 a marine suspension feeder revealed by chemical and isotopic tracers. J. Hazard. 787 Mater. 2014, 279, 32-37. 788 Jeong, J.; Lee, Y.-j.; Hwang, Y. s.; Hong, I. S., Selective detection and 33. 789 quantification of carbon nanotubes in soil. *Environ. Toxicol. Chem.* **2015**, *34*, (9), 790 1969-1974. 791 34. Petersen, E. J.; Henry, T. B., Methodological considerations for testing the 792 ecotoxicity of carbon nanotubes and fullerenes: Review. Environ. Toxicol. Chem. 793 **2012**, *31*, (1), 60-72. 794 35. Su, Y.; Yan, X. M.; Pu, Y. B.; Xiao, F.; Wang, D. S.; Yang, M., Risks of Single-795 Walled Carbon Nanotubes Acting as Contaminants-Carriers: Potential Release of 796 Phenanthrene in Japanese Medaka (Oryzias latipes). Environ. Sci. Technol. 2013, 797 47, (9), 4704-4710. 798 36. Chilek, J. L.; Wang, R. H.; Draper, R. K.; Pantano, P., Use of Gel 799 Electrophoresis and Raman Spectroscopy to Characterize the Effect of the 800 Electronic Structure of Single-Walled Carbon Nanotubes on Cellular Uptake. Anal. 801 Chem. 2014, 86, (6), 2882-2887. 802 37. Wang, R. H.; Mikoryak, C.; Chen, E.; Li, S.; Pantano, P.; Draper, R. K., Gel Electrophoresis Method to Measure the Concentration of Single-Walled Carbon 803 804 Nanotubes Extracted from Biological Tissue. Anal. Chem. 2009, 81, (8), 2944-805 2952. 806 38. Tong, L.; Liu, Y. X.; Dolash, B. D.; Jung, Y.; Slipchenko, M. N.; Bergstrom, D. E.; Cheng, J. X., Label-free imaging of semiconducting and metallic carbon 807

808 nanotubes in cells and mice using transient absorption microscopy. Nat. 809 Nanotech. 2012, 7, (1), 56-61. 810 von der Kammer, F.; Ferguson, P. L.; Holden, P. A.; Masion, A.; Rogers, K. 39. 811 R.; Klaine, S. J.; Koelmans, A. A.; Horne, N.; Unrine, J. M., Analysis of engineered 812 nanomaterials in complex matrices (environment and biota): General 813 considerations and conceptual case studies. Environ. Toxicol. Chem. 2012, 31, (1), 32-49. 814 815 40. Herrero-Latorre, C.; Alvarez-Mendez, J.; Barciela-Garcia, J.; Garcia-Martin, 816 S.; Pena-Crecente, R. M., Characterization of carbon nanotubes and analytical 817 methods for their determination in environmental and biological samples: A review. Anal. Chem. Acta 2015, 853, 77-94. 818 819 Fagan, J. A.; Bauer, B. J.; Hobbie, E. K.; Becker, M. L.; Walker, A. R. H.; 41. 820 Simpson, J. R.; Chun, J.; Obrzut, J.; Bajpai, V.; Phelan, F. R.; Simien, D.; Huh, J. Y.; 821 Migler, K. B., Carbon Nanotubes: Measuring Dispersion and Length. Adv. Mat. 822 **2011**, *23*, (3), 338-348. 823 42. Decker, J. E.; Walker, A. R. H.; Bosnick, K.; Clifford, C. A.; Dai, L.; Fagan, J.; 824 Hooker, S.; Jakubek, Z. J.; Kingston, C.; Makar, J.; Mansfield, E.; Postek, M. T.; 825 Simard, B.; Sturgeon, R.; Wise, S.; Vladar, A. E.; Yang, L.; Zeisler, R., Sample 826 preparation protocols for realization of reproducible characterization of single-827 wall carbon nanotubes. *Metrologia* **2009**, *46*, (6), 682-692. 828 Petersen, E. J.; Henry, T. B.; Zhao, J.; MacCuspie, R. I.; Kirschling, T. L.; 43. 829 Dobrovolskaia, M. A.; Hackley, V.; Xing, B.; White, J. C., Identification and 830 Avoidance of Potential Artifacts and Misinterpretations in Nanomaterial 831 Ecotoxicity Measurements. Environ. Sci. Technol. 2014, 48, (8), 4226-4246. 832 Worle-Knirsch, J. M.; Pulskamp, K.; Krug, H. F., Oops they did it again! 44. 833 Carbon nanotubes hoax scientists in viability assays. Nano Lett. 2006, 6, (6), 834 1261-1268. 835 45. Jakubek, L. M.; Marangoudakis, S.; Raingo, J.; Liu, X. Y.; Lipscombe, D.; Hurt, 836 R. H., The inhibition of neuronal calcium ion channels by trace levels of yttrium 837 released from carbon nanotubes. *Biomaterials* **2009**, *30*, (31), 6351-6357. 838 Gogos, A.; Kaegi, R.; Zenobi, R.; Bucheli, T. D., Capabilities of asymmetric 46. 839 flow field-flow fractionation coupled to multi-angle light scattering to detect carbon nanotubes in soot and soil. Environmental Science-Nano 2014, 1, (6), 584-840 841 594. 842 47. Tagmatarchis, N.; Zattoni, A.; Reschiglian, P.; Prato, M., Separation and 843 purification of functionalised water-soluble multi-walled carbon nanotubes by 844 flow field-flow fractionation. Carbon 2005, 43, (9), 1984-1989. 845 48. Niyogi, S.; Hu, H.; Hamon, M. A.; Bhowmik, P.; Zhao, B.; Rozenzhak, S. M.; 846 Chen, J.; Itkis, M. E.; Meier, M. S.; Haddon, R. C., Chromatographic Purification of 847 Soluble Single-Walled Carbon Nanotubes (s-SWNTs). J. Am. Chem. Soc. 2001, 123, 848 (4), 733-734. 849 49. Fagan, J. A.; Khripin, C. Y.; Batista, C. A. S.; Simpson, J. R.; Haroz, E. H.; 850 Walker, A. R. H.; Zheng, M., Isolation of Specific Small-Diameter Single-Wall 851 Carbon Nanotube Species via Aqueous Two-Phase Extraction. Adv. Mat. 2014, 26, 852 (18), 2800-2804. 853 50. Pycke, B. F. G.; Benn, T. M.; Herckes, P.; Westerhoff, P.; Halden, R. U., 854 Strategies for quantifying C-60 fullerenes in environmental and biological 855 samples and implications for studies in environmental health and ecotoxicology.

856 Trends Anal. Chem. **2011**, *30*, (1), 44-57.

857 Pycke, B. F. G.; Chao, T. C.; Herckes, P.; Westerhoff, P.; Halden, R. U., 51. 858 Beyond nC(60): strategies for identification of transformation products of 859 fullerene oxidation in aquatic and biological samples. Anal. Bioanal. Chem. 2012, 860 404, (9), 2583-2595. 861 Isaacson, C. W.; Kleber, M.; Field, J. A., Quantitative analysis of fullerene 52. 862 nanomaterials in environmental systems: A critical review. Environ. Sci. Technol. 863 **2009**, *43*, (17), 6463-6474. Wang, J. F.; Wages, M.; Yu, S. Y.; Maul, J. D.; Mayer, G.; Hope-Weeks, L.; 864 53. 865 Cobb, G. P., Bioaccumulation of fullerene (C₆₀) and corresponding catalase elevation in Lumbriculus variegatus. Environ. Toxicol. Chem. 2014, 33, (5), 1135-866 867 1141. Pakarinen, K.; Petersen, E. J.; Leppanen, M. T.; Akkanen, J.; Kukkonen, J. V. 868 54. 869 K., Adverse effects of fullerenes (nC(60)) spiked to sediments on Lumbriculus 870 variegatus (Oligochaeta). Environ. Pollut. 2011, 159, (12), 3750-3756. 871 55. Tervonen, K.; Waissi, G.; Petersen, E. J.; Akkanen, J.; Kukkonen, J. V. K., 872 Analysis of fullerene-C₆₀ and kinetic measurements for its accumulation and 873 depuration in *Daphnia magna*. *Environ*. *Toxicol*. *Chem*. **2010**, *29*, (5), 1072-1078. 874 56. Pakarinen, K.; Petersen, E. J.; Alvila, L.; Waissi-Leinonen, G. C.; Akkanen, J.; 875 Leppänen, M. T.; Kukkonen, J. V. K., A screening study on the fate of fullerenes 876 (nC60) and their toxic implications in natural freshwaters. Environ. Toxicol. 877 Chem. 2013, 32, (6), 1224-1232. 878 57. Petersen, E. J.; Lam, T.; Gorham, J. M.; Scott, K. C.; Long, C. J.; Stanley, D.; 879 Sharma, R.; Liddle, J. A.; Pellegrin, B.; Nguyen, T., Methods to assess the impact of 880 UV irradiation on the surface chemistry and structure of multiwall carbon 881 nanotube epoxy nanocomposites. Carbon 2014, 69, 194-205. 882 58. Nowack, B.; David, R. M.; Fissan, H.; Morris, H.; Shatkin, J. A.; Stintz, M.; 883 Zepp, R.; Brouwer, D., Potential release scenarios for carbon nanotubes used in 884 composites. Environ. Intl. 2013, 59, 1-11. 885 Kingston, C.; Zepp, R.; Andrady, A.; Boverhof, D.; Fehir, R.; Hawkins, D.; 59. 886 Roberts, J.; Savre, P.; Shelton, B.; Sultan, Y.; Vejins, V.; Wohlleben, W., Release 887 characteristics of selected carbon nanotube polymer composites. Carbon 2014, 888 68, 33-57. 889 60. Wohlleben, W.; Meier, M. W.; Vogel, S.; Landsiedel, R.; Cox, G.; Hirth, S.; 890 Tomovic, Z., Elastic CNT-polyurethane nanocomposite: synthesis, performance 891 and assessment of fragments released during use. Nanoscale 2013, 5, (1), 369-892 380. 893 61. Green, J. M.; Irin, F.; Canas, J. E.; Saed, M. A. Detection of Carbon Nanotubes 894 by Microwave-Induced Heating. 2012. 895 Irin, F.; Shrestha, B.; Canas, J. E.; Saed, M. A.; Green, M. J., Detection of 62. 896 carbon nanotubes in biological samples through microwave-induced heating. 897 Carbon 2012, 50, (12), 4441-4449. 898 63. Petersen, E. J.; Diamond, S. A.; Kennedy, A. J.; Goss, G. G.; Ho, K.; Lead, J.; 899 Hanna, S. K.; Hartmann, N. B.; Hund-Rinke, K.; Mader, B.; Manier, N.; Pandard, P.; 900 Salinas, E. R.; Sayre, P., Adapting OECD Aquatic Toxicity Tests for Use with 901 Manufactured Nanomaterials: Key Issues and Consensus Recommendations. 902 Environ. Sci. Technol. 2015, 49, (16), 9532-9547. 903 64. Li, Z. F.; Luo, G. H.; Zhou, W. P.; Wei, F.; Xiang, R.; Liu, Y. P., The quantitative 904 characterization of the concentration and dispersion of multi-walled carbon

- 905 nanotubes in suspension by spectrophotometry. Nanotechnol. 2006, 17, (15), 906 3692-3698. 907 65. Cerrillo, C.; Barandika, G.; Igartua, A.; Areitioaurtena, O.; Marcaide, A.; 908 Mendoza, G., Ecotoxicity of multiwalled carbon nanotubes: Standardization of the 909 dispersion methods and concentration measurements. Environ. Toxicol. Chem. 910 2015, 34, (8), 1854-1862. 911 Porter, A. E.; Gass, M.; Muller, K.; Skepper, J. N.; Midgley, P. A.; Welland, M., 66. 912 Direct imaging of single-walled carbon nanotubes in cells. Nat. Nanotech. 2007, 913 2, (11), 713-717. 914 Mavrocordatos, D.; Perret, D., Non-artifacted specimen preparation for 67. 915 transmission electron microscopy of submicron soil particles. Comm. Soil Sci. 916 Plant Anal. 1995, 26, (15-16), 2593-2602. 917 Gogos, A. Engineered nanomaterials in the agricultural environment: 68. 918 current state of applications and development of analytical methods. Ph. D. 919 Thesis No. 22589. Swiss Federal Insitute of Technology (ETH), 2015. 920 Prasad, A.; Lead, J. R.; Baalousha, M., An electron microscopy based 69. 921 method for the detection and quantification of nanomaterial number 922 concentration in environmentally relevant media. Sci. Tot. Environ. 2015, 537, 923 479-486. 924 Gottschalk, F.; Sun, T. Y.; Nowack, B., Environmental concentrations of 70. 925 engineered nanomaterials: Review of modeling and analytical studies. Environ. 926 Pollut. 2013, 181, 287-300. 927 Schwab, F.; Bucheli, T. D.; Lukhele, L. P.; Magrez, A.; Nowack, B.; Sigg, L.; 71. 928 Knauer, K., Are Carbon Nanotube Effects on Green Algae Caused by Shading and 929 Agglomeration? Environ. Sci. Technol. 2011, 45, (14), 6136-6144. 930 72. Pohl, K.; Cantwell, M.; Herckes, P.; Lohmann, R., Black carbon 931 concentrations and sources in the marine boundary layer of the tropical Atlantic 932 Ocean using four methodologies. Atmos. Chem. Phys. 2014, 14, (14), 7431-7443. 933 Dahm, M. M.; Evans, D. E.; Schubauer-Berigan, M. K.; Birch, M. E.; Deddens, 73. 934 J. A., Occupational Exposure Assessment in Carbon Nanotube and Nanofiber 935 Primary and Secondary Manufacturers: Mobile Direct-Reading Sampling. Ann. 936 Occup. Hyg. 2013, 57, (3), 328-344. 937 74. Dahm, M. M.; Schubauer-Berigan, M. K.; Evans, D. E.; Birch, M. E.; 938 Fernback, J. E.; Deddens, J. A., Carbon Nanotube and Nanofiber Exposure 939 Assessments: An Analysis of 14 Site Visits. Ann. Occup. Hyg. 2015, 59, (6), 705-940 723. 941 75. Vo, E.; Zhuang, Z. Q.; Birch, E.; Zhao, Q.; Horvatin, M.; Liu, Y. W., 942 Measurement of Mass-Based Carbon Nanotube Penetration through Filtering 943 Facepiece Respirator Filtering Media. Ann. Occup. Hyg. 2014, 58, (5), 646-656. 944 76. Hashimoto, N.; Ogura, I.; Kotake, M.; Kishimoto, A.; Honda, K., Evaluating 945 the capabilities of portable black carbon monitors and photometers for 946 measuring airborne carbon nanotubes. J. Nano. Res. 2013, 15, (11), Article 947 Number: UNSP 2033 948 Zhang, L. W.; Petersen, E. J.; Habteselassie, M. Y.; Mao, L.; Huang, Q. G., 77. 949 Degradation of multiwall carbon nanotubes by bacteria. *Environ. Pollut.* **2013**, 181, 335-339. 950 951 78. Flores-Cervantes, D. X.; Maes, H. M.; Schaffer, A.; Hollender, J.; Kohler, H. 952 P., Slow biotransformation of carbon nanotubes by horseradish peroxidase.
- 953 Environ. Sci. Technol. **2014**, 48, (9), 4826-34.

- 954 79. Andon, F. T.; Kapralov, A. A.; Yanamala, N.; Feng, W. H.; Baygan, A.; 955 Chambers, B. J.; Hultenby, K.; Ye, F.; Toprak, M. S.; Brandner, B. D.; Fornara, A.; 956 Klein-Seetharaman, J.; Kotchey, G. P.; Star, A.; Shvedova, A. A.; Fadeel, B.; Kagan, 957 V. E., Biodegradation of Single-Walled Carbon Nanotubes by Eosinophil 958 Peroxidase. Small 2013, 9, (16), 2721-2729. 959 Farrera, C.; Bhattacharya, K.; Lazzaretto, B.; Andon, F. T.; Hultenby, K.; 80. 960 Kotchey, G. P.; Star, A.; Fadeel, B., Extracellular entrapment and degradation of 961 single-walled carbon nanotubes. *Nanoscale* **2014**, *6*, (12), 6974-6983. 962 81. Kagan, V. E.; Kapralov, A. A.; St Croix, C. M.; Watkins, S. C.; Kisin, E. R.; 963 Kotchey, G. P.; Balasubramanian, K.; Vlasova, II; Yu, J.; Kim, K.; Seo, W.; 964 Mallampalli, R. K.; Star, A.; Shvedova, A. A., Lung Macrophages "Digest" Carbon 965 Nanotubes Using a Superoxide/Peroxynitrite Oxidative Pathway. ACS Nano 966 **2014**, *8*, (6), 5610-5621. 967 Kotchey, G. P.; Hasan, S. A.; Kapralov, A. A.; Ha, S. H.; Kim, K.; Shvedova, A. 82. A.; Kagan, V. E.; Star, A., A Natural Vanishing Act: The Enzyme-Catalyzed 968 Degradation of Carbon Nanomaterials. Acc. Chem. Res. 2012, 45, (10), 1770-969 970 1781. Kotchey, G. P.; Zhao, Y.; Kagan, V. E.; Star, A., Peroxidase-mediated 971 83. 972 biodegradation of carbon nanotubes in vitro and in vivo. Adv. Drug Deliv. Rev. 973 2013, 65, (15), 1921-1932. 974 Gottschalk, F.; Sonderer, T.; Scholz, R. W.; Nowack, B., Modeled 84. 975 environmental concentrations of engineered nanomaterials (TiO₂, ZnO, Ag, CNT, 976 Fullerenes) for different regions. Environ. Sci. Technol. 2009, 43, (24), 9216-977 9222. 978 85. Mueller, N. C.; Nowack, B., Exposure modeling of engineered nanoparticles 979 in the environment. Environ. Sci. Technol. 2008, 42, (12), 4447-4453. 980 Sun, T. Y.; Conroy, G.; Donner, E.; Hungerbuhler, K.; Lombi, E.; Nowack, B., 86. 981 Probabilistic modelling of engineered nanomaterial emissions to the 982 environment: a spatio-temporal approach. Environmental Science-Nano 2015, 2, 983 (4), 340-351. 984 87. Montano, M. D.; Badiei, H. R.; Bazargan, S.; Ranville, J. F., Improvements in 985 the detection and characterization of engineered nanoparticles using spICP-MS with microsecond dwell times. Environ. Sci.: Nano 2014, 1, (4), 338-346. 986 987 Borovinskaya, O.; Hattendorf, B.; Tanner, M.; Gschwind, S.; Gunther, D., A 88. prototype of a new inductively coupled plasma time-of-flight mass spectrometer 988 providing temporally resolved, multi-element detection of short signals 989 990 generated by single particles and droplets. J. Anal. At. Spectrom. 2013, 28, (2), 991 226-233. 992 Borovinskaya, O.; Gschwind, S.; Hattendorf, B.; Tanner, M.; Gunther, D., 89. 993 Simultaneous Mass Quantification of Nanoparticles of Different Composition in a 994 Mixture by Microdroplet Generator-ICPTOFMS. Anal. Chem. 2014, 86, (16), 8142-995 8148. 996 90. Zhu, Y.; Ran, T. C.; Li, Y. G.; Guo, J. X.; Li, W. X., Dependence of the 997 cytotoxicity of multi-walled carbon nanotubes on the culture medium. 998 Nanotechnol. 2006, 17, (18), 4668-4674. 999 91. Guo, X.; Dong, S.; Petersen, E. J.; Gao, S.; Huang, Q.; Mao, L., Biological 1000 Uptake and Depuration of Radio-labeled Graphene by Daphnia magna. Environ.
- 1001 *Sci. Technol.* **2013,** *47*, (21), 12524-12531.

1002 92. Li, D.; Fortner, J. D.; Johnson, D. R.; Chen, C.; Li, Q. L.; Alvarez, P. J. J., 1003 Bioaccumulation of ¹⁴C₆₀ by the Earthworm *Eisenia fetida*. *Environ. Sci. Technol.* 1004 2010, 44, (23), 9170-9175. 1005 93. Liu, J. H.; Yang, S. T.; Wang, X.; Wang, H. F.; Liu, Y. M.; Luo, P. J. G.; Liu, Y. F.; 1006 Sun, Y. P., Carbon Nanoparticles Trapped in Vivo-Similar to Carbon Nanotubes in 1007 Time-Dependent Biodistribution. ACS Appl. Mater. Interf. 2014, 6, (16), 14672-1008 14678. 1009 94. Feng, Y. P.; Lu, K.; Mao, L.; Guo, X. K.; Gao, S. X.; Petersen, E. J., Degradation 1010 of C-14-labeled few layer graphene via Fenton reaction: Reaction rates, 1011 characterization of reaction products, and potential ecological effects. Water Res. 1012 2015, 84, 49-57. 1013 Mao, L.; Hu, M.; Pan, B.; Xie, Y.; Petersen, E. J., Biodistribution and toxicity 95. of radio-labeled few layer graphene in mice after intratracheal instillation. Part. 1014 1015 *Fibre Toxicol.* **2016**, *13*, (1), 1-12. Wang, H. F.; Wang, J.; Deng, X. Y.; Sun, H. F.; Shi, Z. J.; Gu, Z. N.; Liu, Y. F.; 1016 96. 1017 Zhao, Y. L., Biodistribution of carbon single-wall carbon nanotubes in mice. J. 1018 Nanosci. Nanotechnol. 2004, 4, (8), 1019-1024. 1019 97. Wang, H. F.; Yang, S. T.; Cao, A. N.; Liu, Y. F., Quantification of Carbon 1020 Nanomaterials in Vivo. Acc. Chem. Res. 2013, 46, (3), 750-760. 1021 Singh, R.; Pantarotto, D.; Lacerda, L.; Pastorin, G.; Klumpp, C.; Prato, M.; 98. 1022 Bianco, A.; Kostarelos, K., Tissue biodistribution and blood clearance rates of 1023 intravenously administered carbon nanotube radiotracers. Proc. Natl. Acad. Sci. 1024 U.S.A. 2006, 103, (9), 3357-3362. 1025 Zhang, L. W.; Petersen, E. J.; Zhang, W.; Chen, Y. S.; Cabrera, M.; Huang, Q. 99. 1026 G., Interactions of C-14-labeled multi-walled carbon nanotubes with soil minerals 1027 in water. Environ. Pollut. 2012, 166, 75-81. 1028 Zhao, Q.; Petersen, E. J.; Cornelis, G.; Wang, X.; Guo, X.; Tao, S.; Xing, B., 100. 1029 Retention of 14C-labeled multiwall carbon nanotubes by humic acid and 1030 polymers: Roles of macromolecule properties. Carbon 2016, 99, 229-237. 1031 Zhang, D.; Pan, B.; Cook, R. L.; Xing, B. S., Multi-walled carbon nanotube 101. 1032 dispersion by the adsorbed humic acids with different chemical structures. 1033 Environ. Pollut. 2015, 196, 292-299. Hyung, H.; Kim, J. H., Natural organic matter (NOM) adsorption to multi-1034 102. 1035 walled carbon nanotubes: Effect of NOM characteristics and water quality 1036 parameters. Environ. Sci. Technol. 2008, 42, (12), 4416-4421. Parks, A. N.; Chandler, G. T.; Ho, K. T.; Burgess, R. M.; Ferguson, P. L., 1037 103. 1038 Environmental biodegradability of [C¹⁴] single-walled carbon nanotubes by 1039 Trametes versicolor and natural microbial cultures found in New Bedford Harbor 1040 sediment and aerated wastewater treatment plant sludge. Environ. Toxicol. Chem. 1041 **2015,** *34*, (2), 247-251. 1042 Bitter, J. L.; Yang, J.; Beigzadeh Milani, S.; Jafvert, C. T.; Fairbrother, D. H., 104. 1043 Transformations of oxidized multiwalled carbon nanotubes exposed to UVC (254 1044 nm) irradiation. Environmental Science: Nano 2014, 1, (4), 324-337. 1045 Hou, W.-C.; BeigzadehMilani, S.; Jafvert, C. T.; Zepp, R. G., Photoreactivity 105. of Unfunctionalized Single-Wall Carbon Nanotubes Involving Hydroxyl Radical: 1046 1047 Chiral Dependency and Surface Coating Effect. Environmental Science & 1048 Technology 2014, 48, (7), 3875-3882. Hou, W. C.; BeigzadehMilani, S.; Jafvert, C. T.; Zepp, R. G., Photoreactivity of 1049 106. Unfunctionalized Single-Wall Carbon Nanotubes Involving Hydroxyl Radical: 1050

1051 Chiral Dependency and Surface Coating Effect. Environ. Sci. Technol. 2014, 48, 1052 (7), 3875-3882. Bitter, J. L.; Yang, J.; Milani, S. B.; Jafvert, C. T.; Fairbrother, D. H., 1053 107. 1054 Transformations of oxidized multiwalled carbon nanotubes exposed to UVC (254 1055 nm) irradiation. Environmental Science-Nano 2014, 1, (4), 324-337. 1056 Chen, C. Y.; Jafvert, C. T., Photoreactivity of Carboxylated Single-Walled 108. 1057 Carbon Nanotubes in Sunlight: Reactive Oxygen Species Production in Water. 1058 Environ. Sci. Technol. 2010, 44, (17), 6674-6679. 1059 109. Khripin, C. Y.; Tu, X.; Heddleston, J. M.; Silvera-Batista, C.; Hight Walker, A. 1060 R.; Fagan, J.; Zheng, M., High-Resolution Length Fractionation of Surfactant-Dispersed Carbon Nanotubes. Anal. Chem. 2013, 85, (3), 1382-1388. 1061 1062 Hull, M.; Kennedy, A. J.; Detzel, C.; Vikesland, P.; Chappell, M. A., Moving 110. 1063 beyond Mass: The Unmet Need to Consider Dose Metrics in Environmental 1064 Nanotoxicology Studies. Environ. Sci. Technol. 2012, 46, (20), 10881-10882. 1065 111. Roebben, G.; Rasmussen, K.; Kestens, V.; Linsinger, T. P. J.; Rauscher, H.; 1066 Emons. H.: Stamm. H., Reference materials and representative test materials: the 1067 nanotechnology case. J. Nano. Res. 2013, 15, (3), Article Number: 1455 ISO (International Organization for Standardization), TS 10798: 1068 112. 1069 Nanotechnologies -- Charaterization of single-wall carbon nanotubes using 1070 scanning electron microscopy and energy dispersive X-ray spectrometry 1071 analysis. Geneva, Switzerland, 2011. 113. ISO (International Organization for Standardization), TS 10867: 1072 1073 Nanotechnologies -- Characterization of single-wall carbon nanotubes using near 1074 infrared photoluminescence spectroscopy. Geneva, Switzerland, 2010. 1075 ISO (International Organization for Standardization), TS 10868: 114. 1076 Nanotechnologies -- Characterization of single-wall carbon nanotubes using 1077 ultraviolet-visible-near infrared (UV-Vis-NIR) absorption spectroscopy. Geneva, 1078 Switzerland: 2011. 1079 ISO (International Organization for Standardization), TS 10929: 115. 1080 Nanotechnologies -- Characterization of multiwall carbon nanotube (MWCNT) 1081 samples. Geneva, Switzerland, 2012. 1082 ISO (International Organization for Standardization), TS 11251: 116. Characterization of volatile components in single-wall carbon nanotube samples 1083 using evolved gas analysis/gas chromatograph-mass spectrometry. Geneva, 1084 1085 Switzerland, 2010. 1086 ISO (International Organization for Standardization), TS 11888: 117. 1087 Nanotechnologies -- Characterization of multiwall carbon nanotubes --Mesoscopic shape factors. Geneva, Switzerland, 2011. 1088 ISO (International Organization for Standardization), TS 13278: 1089 118. Nanotechnologies -- Determination of elemental impurities in samples of carbon 1090 1091 nanotubes using inductively coupled plasma mass spectrometry. Geneva. Switzerland, 2011. 1092 1093 119. Organization for Economic Cooperation and Development. 2012. 1094 Bioaccumulation in fish: Aqueous and Dietary Exposure. OECD Guideline 305. 1095 Paris. F. 1096 120. Edgington, A. J.; Roberts, A. P.; Taylor, L. M.; Alloy, M. M.; Reppert, J.; Rao, 1097 A. M.; Ma, J. D.; Klaine, S. J., The influence of natural organic matter on the toxicity 1098 of multiwalled carbon nanotubes. Environ. Toxicol. Chem. 2010, 29, (11), 2511-1099 2518.

1100 Leeuw, T. K.; Reith, R. M.; Simonette, R. A.; Harden, M. E.; Cherukuri, P.; 121. 1101 Tsyboulski, D. A.; Beckingham, K. M.; Weisman, R. B., Single-walled carbon 1102 nanotubes in the intact organism: Near-IR imaging and biocompatibility studies 1103 in Drosophila. Nano Lett. 2007, 7, (9), 2650-2654. 1104 Mouchet, F.; Landois, P.; Puech, P.; Pinelli, E.; Flahaut, E.; Gauthier, L., 122. 1105 Carbon nanotube ecotoxicity in amphibians: assessment of multiwalled carbon nanotubes and comparison with double-walled carbon nanotubes. Nanomedicine 1106 1107 **2010**, *5*, (6), 963-974. 1108 123. Mouchet, F.; Landois, P.; Sarremejean, E.; Bernard, G.; Puech, P.; Pinelli, E.; 1109 Flahaut, E.; Gauthier, L., Characterisation and in vivo ecotoxicity evaluation of double-wall carbon nanotubes in larvae of the amphibian Xenopus laevis. Aquat. 1110 1111 Toxicol. 2008, 87, (2), 127-137. 1112 124. Maes, H. M.; Stibany, F.; Giefers, S.; Daniels, B.; Deutschmann, B.; 1113 Baumgartner, W.; Schaffer, A., Accumulation and Distribution of Multiwalled 1114 Carbon Nanotubes in Zebrafish (Danio rerio). Environ. Sci. Technol. 2014, 48, 1115 (20), 12256-12264. 1116 125. Li, S. B.; Irin, F.; Atore, F. O.; Green, M. J.; Canas-Carrell, J. E., Determination 1117 of multi-walled carbon nanotube bioaccumulation in earthworms measured by a 1118 microwave-based detection technique. Sci. Tot. Environ. 2013, 445, 9-13. 1119 Rösslein, M.; Elliott, J. T.; Salit, M. L.; Petersen, E. J.; Hirsch, C.; Krug, H. F.; 126. 1120 Wick, P., The use of cause-and-effect analysis to design a high quality nano-1121 cytotoxicology assay. Chem. Res. Toxicol. 2014, 27, (10), 1877-1884. 1122 Hammes, K.; Schmidt, M. W. I.; Smernik, R. J.; Currie, L. A.; Ball, W. P.; 127. 1123 Nguyen, T. H.; Louchouarn, P.; Houel, S.; Gustafsson, Ö.; Elmquist, M.; Cornelissen, 1124 G.; Skjemstad, J. O.; Masiello, C. A.; Song, J.; Peng, P. a.; Mitra, S.; Dunn, J. C.; 1125 Hatcher, P. G.; Hockaday, W. C.; Smith, D. M.; Hartkopf-Fröder, C.; Böhmer, A.; 1126 Lüer, B.; Huebert, B. J.; Amelung, W.; Brodowski, S.; Huang, L.; Zhang, W.; 1127 Gschwend, P. M.; Flores-Cervantes, D. X.; Largeau, C.; Rouzaud, J.-N.; Rumpel, C.; 1128 Guggenberger, G.; Kaiser, K.; Rodionov, A.; Gonzalez-Vila, F. J.; Gonzalez-Perez, J. 1129 A.; de la Rosa, J. M.; Manning, D. A. C.; López-Capél, E.; Ding, L., Comparison of 1130 quantification methods to measure fire-derived (black/elemental) carbon in 1131 soils and sediments using reference materials from soil, water, sediment and the 1132 atmosphere. *Global Biogeochem. Cycles* **2007**, *21*, (3), Article Number: GB3016 1133 128. Cleveland, D.; Long, S. E.; Pennington, P. L.; Cooper, E.; Fulton, M. H.; Scott, 1134 G. I.; Brewer, T.; Davis, J.; Petersen, E. J.; Wood, L., Pilot estuarine mesocosm 1135 study on the environmental fate of silver nanomaterials leached from consumer products. Sci. Tot. Environ. 2012, 421, 267-272. 1136 1137 129. Bour, A.; Mouchet, F.; Silvestre, J.; Gauthier, L.; Pinelli, E., Environmentally 1138 relevant approaches to assess nanoparticles ecotoxicity: A review. J. Hazard. 1139 Mater. 2015, 283, 764-777. 130. Chen, B. L.; Selegue, J. P., Separation and characterization of single-walled 1140 1141 and multiwalled carbon nanotubes by using flow field-flow fractionation. Anal. 1142 *Chem.* **2002**, *74*, (18), 4774-4780. 1143 Moon, M. H.; Kang, D. J.; Jung, J. H.; Kim, J. M., Separation of carbon 131. nanotubes by frit inlet asymmetrical flow field-flow fractionation. J. Sep. Sci. 1144 1145 **2004**, *27*, (9), 710-717. 1146 Chun, J.; Fagan, J. A.; Hobbie, E. K.; Bauer, B. J., Size separation of single-132. wall carbon nanotubes by flow-field flow fractionation. Anal. Chem. 2008, 80, (7), 1147 1148 2514-2523.

1149 Gigault, J.; Grassl, B.; Lespes, G., Size characterization of the associations 133. 1150 between carbon nanotubes and humic acids in aqueous media by asymmetrical 1151 flow field-flow fractionation combined with multi-angle light scattering. 1152 Chemosphere 2012, 86, (2), 177-182. 1153 Doorn, S. K.; Fields, R. E.; Hu, H.; Hamon, M. A.; Haddon, R. C.; Selegue, J. P.; 134. 1154 Majidi, V., High resolution capillary electrophoresis of carbon nanotubes. *JACS* 1155 2002, 124, (12), 3169-3174. 1156 Doorn, S. K.; Strano, M. S.; O'Connell, M. J.; Haroz, E. H.; Rialon, K. L.; Hauge, 135. 1157 R. H.; Smalley, R. E., Capillary electrophoresis separations of bundled and 1158 individual carbon nanotubes. J. Phys. Chem. B 2003, 107, (25), 6063-6069. Suarez, B.; Simonet, B. M.; Cardenas, S.; Valcarcel, M., Separation of carbon 1159 136. 1160 nanotubes in aqueous medium by capillary electrophoresis. J. Chromat. A 2006, 1161 1128, (1-2), 282-289. 1162 137. Lopez-Pastor, M.; Dominguez-Vidal, A.; Ayora-Canada, M. J.; Simonet, B. M.; Lendl, B.; Valcarcel, M., Separation of single-walled carbon nanotubes by use 1163 1164 of ionic liquid-aided capillary electrophoresis. Anal. Chem. 2008, 80, (8), 2672-1165 2679. 1166 138. Fagan, J. A.; Becker, M. L.; Chun, J. H.; Nie, P. T.; Bauer, B. J.; Simpson, J. R.; 1167 Hight-Walker, A.; Hobbie, E. K., Centrifugal length separation of carbon 1168 nanotubes. Langmuir 2008, 24, (24), 13880-13889. 1169 Komatsu, N.; Wang, F., A Comprehensive Review on Separation Methods 139. 1170 and Techniques for Single-Walled Carbon Nanotubes. *Materials* **2010**, *3*, (7), 1171 3818-3844. 1172 Fagan, J. A.; Becker, M. L.; Chun, J.; Hobbie, E. K., Length fractionation of 140. 1173 carbon nanotubes using centrifugation. Adv. Mat. 2008, 20, (9), 1609-1613. 1174 141. Duesberg, G. S.; Blau, W.; Byrne, H. J.; Muster, J.; Burghard, M.; Roth, S., 1175 Chromatography of carbon nanotubes. *Synthetic Metals* **1999**, *103*, (1-3), 2484-1176 2485. Duesberg, G. S.; Burghard, M.; Muster, J.; Philipp, G.; Roth, S., Separation of 1177 142. 1178 carbon nanotubes by size exclusion chromatography. *Chem. Commun.* **1998**, (3), 1179 435-436. Duesberg, G. S.; Muster, J.; Krstic, V.; Burghard, M.; Roth, S., 1180 143. 1181 Chromatographic size separation of single-wall carbon nanotubes. Appl. Phys. A: Mater Sci. Proces. 1998, 67, (1), 117-119. 1182 Farkas, E.; Anderson, M. E.; Chen, Z. H.; Rinzler, A. G., Length sorting cut 1183 144. 1184 single wall carbon nanotubes by high performance liquid chromatography. Chem. Phys. Let. 2002, 363, (1-2), 111-116. 1185 1186 145. Zhao, B.; Hu, H.; Niyogi, S.; Itkis, M. E.; Hamon, M. A.; Bhowmik, P.; Meier, 1187 M. S.; Haddon, R. C., Chromatographic Purification and Properties of Soluble Single-Walled Carbon Nanotubes. J. Am. Chem. Soc. 2001, 123, (47), 11673-1188 11677. 1189 1190 146. Ziegler, K. J.; Schmidt, D. J.; Rauwald, U.; Shah, K. N.; Flor, E. L.; Hauge, R. H.; Smalley, R. E., Length-dependent extraction of single-walled carbon 1191 nanotubes. Nano Lett. 2005, 5, (12), 2355-2359. 1192 Flavel, B. S.; Moore, K. E.; Pfohl, M.; Kappes, M. M.; Hennrich, F., Separation 1193 147. 1194 of Single-Walled Carbon Nanotubes with a Gel Permeation Chromatography 1195 System. ACS Nano 2014, 8, (2), 1817-1826. Suarez, B.; Moliner-Martinez, Y.; Cardenas, S.; Simonet, B. M.; Valcarcel, M., 1196 148. Monitoring of carboxylic carbon nanotubes in surface water by using 1197

1198 multiwalled carbon nanotube-modified filter as preconcentration unit. Environ. 1199 Sci. Technol. 2008, 42, (16), 6100-6104. 1200 Ziolkowski, L. A.; Druffel, E. R. M., The feasibility of isolation and detection 149. 1201 of fullerenes and carbon nanotubes using the benzene polycarboxylic acid 1202 method. Mar. Poll. Bull. 2009, 59, (4-7), 213-218. 1203 Mattison, N. T.; O'Carroll, D. M.; Rowe, R. K.; Petersen, E. J., Impact of 150. 1204 Porous Media Grain Size on the Transport of Multi-walled Carbon Nanotubes. 1205 Environ. Sci. Technol. 2011, 45, (22), 9765-9775. 1206 151. O'Carroll, D. M.; Liu, X.; Mattison, N. T.; Petersen, E. J., Impact of diameter 1207 on carbon nanotube transport in sand. J. Coll. Interf. Sci. 2013, 390, 96-104. Parks, A. N.; Portis, L. M.; Schierz, P. A.; Washburn, K. M.; Perron, M. M.; 1208 152. 1209 Burgess, R. M.; Ho, K. T.; Chandler, G. T.; Ferguson, P. L., Bioaccumulation and 1210 toxicity of single-walled carbon nanotubes to benthic organisms at the base of 1211 the marine food chain. *Environ. Toxicol. Chem.* **2013**, *32*, (6), 1270-1277. 1212 153. Hennrich, F.; Krupke, R.; Lebedkin, S.; Arnold, K.; Fischer, R.; Resasco, D. E.; Kappes, M., Raman spectroscopy of individual single-walled carbon nanotubes 1213 1214 from various sources. J. Phys. Chem. B 2005, 109, (21), 10567-10573. Heeg, S.; Malic, E.; Casiraghi, C.; Reich, S., Quantitative composition of a 1215 154. 1216 single-walled carbon nanotube sample: Raman scattering versus 1217 photoluminescence. Phys. Status Sol. B: Basic Solid State Phys. 2009, 246, (11-12), 1218 2740-2743. 1219 155. Dresselhaus, M. S.; Jorio, A.; Hofmann, M.; Dresselhaus, G.; Saito, R., 1220 Perspectives on Carbon Nanotubes and Graphene Raman Spectroscopy. Nano 1221 Lett. 2010, 10, (3), 751-758. 1222 Caballero-Diaz, E.; Guzman-Ruiz, R.; Malagon, M. M.; Simonet, B. M.; 156. 1223 Valcarcel, M., Effects of the interaction of single-walled carbon nanotubes with 4-1224 nonylphenol on their in vitro toxicity. J. Hazard. Mater. 2014, 275, 107-115. 1225 Lopez-Lorente, A. I.; Simonet, B. M.; Valcarcel, M., Raman spectroscopic 157. characterization of single walled carbon nanotubes: influence of the sample 1226 1227 aggregation state. Analyst 2014, 139, (1), 290-298. 1228 Lopez-Lorente, A. I.; Simonet, B. M.; Valcarcel, M.; Mizaikoff, B., Bare gold 158. 1229 nanoparticles mediated surface-enhanced Raman spectroscopic determination and quantification of carboxylated single-walled carbon nanotubes. Anal. Chem. 1230 1231 Acta 2013, 788, 122-128. Lopez-Lorente, A. I.; Simonet, B. M.; Valcarcel, M., Qualitative detection 1232 159. 1233 and quantitative determination of single-walled carbon nanotubes in mixtures of 1234 carbon nanotubes with a portable Raman spectrometer. Analyst **2013**, 138, (8), 1235 2378-2385. 1236 Marches, R.; Mikoryak, C.; Wang, R. H.; Pantano, P.; Draper, R. K.; Vitetta, E. 160. S., The importance of cellular internalization of antibody-targeted carbon 1237 1238 nanotubes in the photothermal ablation of breast cancer cells. Nanotechnol. 1239 **2011**, 22, (9), Article Number: 095101. 1240 Liu, Z.; Davis, C.; Cai, W. B.; He, L.; Chen, X. Y.; Dai, H. J., Circulation and 161. 1241 long-term fate of functionalized, biocompatible single-walled carbon nanotubes in mice probed by Raman spectroscopy. Proc. Natl. Acad. Sci. U.S.A. 2008, 105, 1242 1243 (5), 1410-1415. 1244 Salzmann, C. G.; Chu, B. T. T.; Tobias, G.; Llewellyn, S. A.; Green, M. L. H., 162. 1245 Quantitative assessment of carbon nanotube dispersions by Raman spectroscopy. Carbon 2007, 45, (5), 907-912. 1246

1247 Athalin, H.; Lefrant, S. L., A correlated method for quantifying mixed and 163. 1248 dispersed carbon nanotubes: analysis of the Raman band intensities and 1249 evidence of wavenumber shift. J. Raman. Spectr. 2005, 36, (5), 400-408. 1250 Nunes, A.; Bussy, C.; Gherardini, L.; Meneghetti, M.; Herrero, M. A.; Bianco, 164. 1251 A.; Prato, M.; Pizzorusso, T.; Al-Jamal, K. T.; Kostarelos, K., In vivo degradation of 1252 functionalized carbon nanotubes after stereotactic administration in the brain 1253 cortex. Nanomedicine 2012, 7, (10), 1485-1494. 1254 Batista, C. A. S.; Zheng, M.; Khripin, C. Y.; Tu, X. M.; Fagan, J. A., Rod 165. 1255 Hydrodynamics and Length Distributions of Single-Wall Carbon Nanotubes Using 1256 Analytical Ultracentrifugation. Langmuir 2014, 30, (17), 4895-4904. 1257 Badireddy, A. R.; Wiesner, M. R.; Liu, J., Detection, Characterization, and 166. 1258 Abundance of Engineered Nanoparticles in Complex Waters by Hyperspectral 1259 Imagery with Enhanced Darkfield Microscopy. Environ. Sci. Technol. 2012, 46, 1260 (18), 10081 - 10088.Mortimer, M.; Gogos, A.; Bartolome, N.; Kahru, A.; Bucheli, T. D.; 1261 167. Slaveykova, V. I., Potential of Hyperspectral Imaging Microscopy for Semi-1262 1263 quantitative Analysis of Nanoparticle Uptake by Protozoa. Environ. Sci. Technol. 1264 2014, 48, (15), 8760-8767. Berciaud, S.; Cognet, L.; Poulin, P.; Weisman, R. B.; Lounis, B., Absorption 1265 168. spectroscopy of individual single-walled carbon nanotubes. Nano Lett. 2007, 7, 1266 1267 (5), 1203-1207. 1268 169. Nedosekin, D. A.; Shashkov, E. V.; Galanzha, E. I.; Hennings, L.; Zharov, V. 1269 P., Photothermal Multispectral Image Cytometry for Quantitative Histology of 1270 Nanoparticles and Micrometastasis in Intact, Stained and Selectively Burned 1271 Tissues. Cytometry Part A 2010, 77A, (11), 1049-1058. 1272 170. Hong, H.; Chen, F.; Cai, W. B., Pharmacokinetic Issues of Imaging with 1273 Nanoparticles: Focusing on Carbon Nanotubes and Quantum Dots. Mol. Imag. 1274 Biol. 2013, 15, (5), 507-520. 1275 Lehman, J. H.; Terrones, M.; Mansfield, E.; Hurst, K. E.; Meunier, V., 171. 1276 Evaluating the characteristics of multiwall carbon nanotubes. *Carbon* **2011**, 49, 1277 (8), 2581-2602. 1278 Mansfield, E.; Kar, A.; Wang, C. M.; Chiaramonti, A. N., Statistical sampling 172. 1279 of carbon nanotube populations by thermogravimetric analysis. Anal. Bioanal. 1280 Chem. 2013, 405, (25), 8207-8213. 1281 Petersen, E. J.; Huang, Q.; Weber, W. J., Jr., Bioaccumulation of radio-173. labeled carbon nanotubes by Eisenia foetida. Environ. Sci. Technol. 2008, 42, (8), 1282 3090-3095. 1283 1284 174. Petersen, E. J.; Huang, Q.; Weber, W. J., Jr., Ecological uptake and 1285 depuration of carbon nanotubes by Lumbriculus variegatus. Environ. Health 1286 Perspect. 2008, 116, (4), 496-500. Petersen, E. J.; Pinto, R. A.; Zhang, L.; Huang, Q. G.; Landrum, P. F.; 1287 175. Weber, W. J., Effects of polyethyleneimine-mediated functionalization of multi-1288 walled carbon nanotubes on earthworm bioaccumulation and sorption by 1289 1290 soils. Environ. Sci. Technol. 2011, 45, (8), 3718–3724. 1291 176. Zhang, L.; Petersen, E. J.; Qingguo, H., Phase Distribution of 14C-1292 Labeled Multiwalled Carbon Nanotubes in Aqueous Systems Containing 1293 Model Solids: Peat. Environ. Sci. Technol. 2011, 45, (4), 1356-1362.

1294 Petersen, E. J.; Akkanen, J.; Kukkonen, J. V. K.; Weber, W. J., Jr., Biological 177. 1295 Uptake and Depuration of Carbon Nano-tubes by Daphnia magna. Environ. Sci. 1296 Technol. 2009, 43, (8), 2969-2975. 1297 Petersen, E. J.; Huang, Q. G.; Weber, W. J., Jr., Relevance of octanol-178. 1298 water distribution measurements to the potential ecological uptake of multi-1299 walled carbon nanotubes. Environ. Toxicol. Chem. 2010, 29, (5), 1106-1112. 1300 179. Larue, C.; Pinault, M.; Czarny, B.; Georgin, D.; Jaillard, D.; Bendiab, N.; 1301 Mayne-L'Hermite, M.; Taran, F.; Dive, V.; Carriere, M., Quantitative evaluation of 1302 multi-walled carbon nanotube uptake in wheat and rapeseed. J. Hazard. Mater. 1303 **2012**, *227*, 155-163. 1304 Zhang, L.; Petersen, E. J.; Habteselassie, M. Y.; Mao, L.; Huang, Q., 180. 1305 Degradation of multiwall carbon nanotubes by bacteria. *Environ. Pollut.* 2013, 1306 181, 335-339. 1307 Rhiem, S.; Riding, M. J.; Baumgartner, W.; Martin, F. L.; Semple, K. T.; Jones, 181. 1308 K. C.; Schaffer, A.; Maes, H. M., Interactions of multiwalled carbon nanotubes with 1309 algal cells: Quantification of association, visualization of uptake, and 1310 measurement of alterations in the composition of cells. *Environ. Pollut.* **2015**, 1311 196, 431-439. Fagan, J. A.; Zheng, M.; Rastogi, V.; Simpson, J. R.; Khripin, C. Y.; Batista, C. 1312 182. 1313 A. S.; Walker, A. R. H., Analyzing Surfactant Structures on Length and Chirality 1314 Resolved (6,5) Single-Wall Carbon Nanotubes by Analytical Ultracentrifugation. 1315 ACS Nano 2013, 7, (4), 3373-3387. Arnold, M. S.; Suntivich, J.; Stupp, S. I.; Hersam, M. C., Hydrodynamic 1316 183. 1317 Characterization of Surfactant Encapsulated Carbon Nanotubes Using an 1318 Analytical Ultracentrifuge. ACS Nano 2008, 2, (11), 2291-2300. Backes, C.; Karabudak, E.; Schmidt, C. D.; Hauke, F.; Hirsch, A.; Wohlleben, 1319 184. 1320 W., Determination of the Surfactant Density on SWCNTs by Analytical 1321 Ultracentrifugation. Chem- Europ. J. 2010, 16, (44), 13176-13184. 1322 Kim, K. T.; Edgington, A. J.; Klaine, S. J.; Cho, J. W.; Kim, S. D., Influence of 185. 1323 multiwalled carbon nanotubes dispersed in natural organic matter on speciation 1324 and bioavailability of copper. Environ. Sci. Technol. 2009, 43, (23), 8979-8984. 1325 Bourdiol, F.; Dubuc, D.; Grenier, K.; Mouchet, F.; Gauthier, L.; Flahaut, E., 186. 1326 Quantitative detection of carbon nanotubes in biological samples by an original 1327 method based on microwave permittivity measurements. Carbon 2015, 81, 535-1328 545. 1329 Cano, A.; Kohl, K.; Deleon, S.; Payton, P.; Irin, F.; Saed, M.; Shah, S. A.; Green, 187. 1330 M. J.; Cañas-Carrell, J. E., Determination of uptake, accumulation, and stress 1331 effects in corn (Zea mays L.) grown in single-wall carbon nanotube contaminated 1332 soil. *Chemosphere* **2016**, in press. 1333 Menendez, J. A.; Arenillas, A.; Fidalgo, B.; Fernandez, Y.; Zubizarreta, L.; 188. 1334 Calvo, E. G.; Bermudez, J. M., Microwave heating processes involving carbon 1335 materials. Fuel Process. Technol. 2010, 91, (1), 1-8. Garner, K. L.; Suh, S.; Lenihan, H. S.; Keller, A. A., Species Sensitivity 1336 189. 1337 Distributions for Engineered Nanomaterials. *Environ. Sci. Technol.* **2015**, *49*. (9). 1338 5753-5759.

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Technique	Overview	Strengths	Limitations
Field-Flow Fractionation 46,47,130-133	technique based on particle diffusion against a hydrodynamic field in the absence of a stationary phase	dispersed CNTs by length, reduces sample polydispersity, possibility for online/offline coupling with a variety of analytical techniques can yield complementary information	operation/method development, high sample dilution during the analysis, possibility of strong particle-membrane interactions may result in low recoveries, separation less efficient (low number of theoretical plates) than with e.g., capillary electrophoresis
Capillary Electrophoresis ¹³⁴⁻ ¹³⁷	Based on the different electrophoretic mobilities of the species (on the basis of their charge/size ratio) through an electrolyte contained in a fused silica capillary when an electrical field is applied; a suspension of CNTs, usually in a surfactant, is subjected to an electric current	Potential separation of bulk samples of CNTs according to the charge/size ratio, length sorting and separation according to bundled/non-bundled can occur; high theoretical plate number, thus potentially superior resolution power, due to the plug-like flow of the electroosmotic flow	Laborious sample preparation for controlled experiments, several important challenges still remain, including limited sensitivity, non- quantitative recoveries, and reproducibility problems; micellular electrokinetic chromatography cannot be used, as CNTs are too large to reside in the intramicellular region
Centrifugation ¹³⁸	Large suspended particles are removed first on basis of difference in sedimentation velocities	Potential isolation of CNTs from matrix, either in sediment or supernatant	Protocol will depend on CNT and matrix, further separation of the fraction is challenging without disturbing neighboring fractions
Density gradient centrifugation ^{139,140}	Particles will equilibrate to their isopycnic (equal bouyancy point) in a density gradient at sufficiently high applied acceleration	Can enable extraction of specifically modified subpopulations, resolves aggregate states	Low processing quantity, kinetic and transport non-idealities can occur, different aggregation states have different buoyant densities.
Size exclusion chromatography ^{48,1} 41-147	A chromatographic method that separates analytes based on their size and shape by differential exclusion from the pores of the stationary phase; no interactions must exist between CNTs and the stationary phase	Relatively simple and inexpensive, good size separation for SWCNTs within a certain length limit and shape	It has mainly been used for short single-walled carbon nanotubes; it is unclear if this technique can separate larger SWCNTs or MWCNTs; prefiltration might be needed; agglomerates can get trapped within the chromatographic column or the prefilter; well dispersed suspensions are required; only for qualitative analysis; no environmental samples have been tested
Matrix Digestion ²³	Different chemicals or solutions are used to dissolve the matrix (e.g., tissues) to facilitate subsequent analytical techniques	Lowers detection limits and removes potential biases for many techniques	Different approaches will likely need to be developed for each type of matrix (e.g., tissue vs. sediment) and may need to be developed for different types of tissues
Micro- nanofiltration ^{13,148}	Use of micro and nanopore- sized filters to separate analytes based on their size	Very simple and inexpensive; at low CNT concentrations, can treat larger volumes than other techniques	Mainly used for CNT suspensions with very little interferences and at low concentrations to avoid clogging the filters; it is difficult to regenerate the CNT suspension for further characterization/quantification; there might be sample losses and filter interferences
Selective Oxidation ^{14,18,149}	Use of thermal or chemical oxidation to separate more refractory carbon fractions (CNTs) from more labile organic carbon	Allows for a cleaner (and easier) subsequent characterization or quantification	Not very reliable in the presence of interfering material or when the oxidation is not complete (e.g. coals, very rich organic carbon environments); recoveries might vary between different types of CNTs

Sonication with surfactant¹⁵

Use of a surfactant to create a stable CNT suspension that can then be separated from the remaining non CNT material that settles down at a different speed Can extract CNTs with varying surface chemistry from sediment, no special equipment is necessary Recoveries vary among SWCNTs with no recognizable pattern; repeatability varies, surfactants may interfere with quantitation procedure

1340Table 1: Extraction and separation techniques to isolate CNTs from

1341 environmental and biological matrices

Method	Overview	Strengths	Limitations
Spectroscopic			
Absorbance ^{65,109,150,151}	Measures absorbance of aqueous sample; can include ultraviolet, visible, or near infrared wavelengths	Readily available in many environmental laboratories	Interference from other sample components, relatively high detection limit, only applicable for aqueous samples
Near infrared fluorescence (NIRF) ^{15,29,31,152}	A specific emission spectra can be used as an identification tool of SWCNTs; the intensity of the fluorescence signal can be used for quantification of SWCNTs	Quantification/Detection at very low limits of detection	Limited to non-functionalized SWCNTs; semi-conducting SWCNTs but not metallic SWCNTs can be detected
Raman ^{13,122,123,153-164}	Measures radial breathing (SWCNT) , G, D and G' vibrational bands in dry and various solvent suspended samples, tissues	Minimal sample preparation, enables CNT characterization, compatible with <i>in vitro</i> and <i>in vivo</i> samples, can be used with a microscope, low detection limits achieved using resonance Raman conditions	Some matrices may produce interferences, sensitive to laser power, requires calibration for quantitative analysis
Spectrometric			
Inorganic Element Analysis ^{29,32}	Measures trace catalytic metallic elemental impurities intercalated in the CNT structure (Cr, Co, Cu, Fe, Mo, Ni, Y, Zn), analysis of bulk metal content; the applicability of this approach could be impacted by removal of the metal catalysts by purification but catalysts located within the CNTs often remain after purification processes	Multi-elemental capability and extreme sensitivity of ICP-MS allow an accurate and selective determination of metal impurities of CNT in a wide range of matrices at ngL ⁻¹ or sub ngL ⁻¹ levels, the rapid sample throughput of this method is attractive for routine screening	Carbon is generally not detectable with standard ICP-MS methods, quantitative sample dissolution is required prior to analysis; incomplete sample digestion, release of metal ions from the CNTs in the sample matrix, or elemental contamination from the sample digestion steps could lead to an important bias in the bulk metal content determination; the feasibility of using this technique could depend partly on if the metal contents of the CNTs are known <i>a priori</i>
Single particle inductively coupled plasma-mass spectrometry (spICP-MS) ²²	Metal catalyst impurities are used as proxies to detect and quantify CNTs; the applicability of this approach could be impacted by removal of the metal catalysts by purification but catalysts located within the CNTs often remain after purification processes	Potential capability for the size, size distribution, and particle number concentration determination of CNT; high selectivity to differentiate CNT at extremely low concentrations from naturally occurring carbon- containing species (i.e. cells, organic detritus, humic acid); very low detection limit	Size/length estimation requires the invalid assumption that metal content is homogeneous among the CNTs, very small particles cannot be separated from the background, leaching of catalysts in the sample matrix prior to spICP-MS analysis can bias the result, only applicable for aqueous samples; the feasibility of using this technique could depend partly on if the metal contents of the CNTs are known <i>a priori</i>
Microscopic			
Atomic Force Microscopy ^{109,165}	Measure the surface features of a sample by dragging a cantilever over the sample; the length of identifiable tubes can be determined by the movements of the cantilever	Most trusted technique for determining number and length	Deposition bias, measurement bias, and detection errors are all possible in most samples
Hyperspectral Imaging ^{166,167}	Measures reflectance spectra of NPs in a darkfield (visual near infrared /short-wave infrared spectral range), resulting in 2D-optical images with full spectral information that contain a full spectrum (400 nm to 1000 nm or 900 nm to 1700 nm, respectively) in each pixel; CNTs appear bright against a dark background	Easy sample preparation, provides optical (i.e. differentiation between single nanotube and nanotube- agglomerate) and spectral information, allows spatial localization of particles, can provide semi-quantitative information, short-wave infrared spectral range could be applicable for detection of SWCNTs	Currently long analysis times, visual near infrared not specific for CNTs, many potential analysis artifacts

Photoacoustic (PA) ^{24,168-170}	PA measures the acoustic response to the rapid volume change resulting from the absorption of an optical pump beam and the transfer of heat to the surrounding environment	Suitable for detection in liquids such as water and complex media such as plants, minimal sample preparation, can be quantifiable, excellent penetration depth enables samples > 100 μ m, works equally well with metallic and semiconducting SWCNTs and MWCNTs, label free, unaffected by some complex media issues including carbon-on-carbon	Signal is dependent on absorption and heat transfer to material surrounding the CNTs, can be 10x lower sensitivity than PT, medium surrounding CNTs must be transparent to the beams, heating laser must overlap with absorbance of the CNTs, signal scales with size of CNT cluster, non-transparent media may cause detection issues, quantification may require diameter and length distributions
Photothermal (PT) ^{24,168,169}	PT measures the optical scattering response of a probe beam to the change in local environment refractive index that results from the absorption of an optical pump beam and the transfer of heat to the surrounding environment	Suitable for detection in liquids such as water and complex media such as plants, minimal sample preparation, can be quantifiable, penetration depth can handle samples up to 10μ m, works equally well with metallic and semiconducting SWCNTs and MWCNTs, label free, unaffected by some complex media issues including carbon-on-carbon, sensitivity down to single particle sensitivity, lower LOD than absorbance-based measurements	Same as Photoacoustic plus is limited to thin samples (< 100 μ m)
Scanning Electron Microscopy and Scanning Transmission Electron Microscopy	Measures the interaction of a finely focused electron beam with the CNTs; secondary electrons, and transmitted electrons can be used for image formation	Provides detailed morphological properties (length, width, shape) of individual CNTs; individual CNTs can be localized in complex matrices based on morphological criteria	Labor intensive, often only qualitative information
Transmission Electron Microscopy (TEM) ^{27,66}	Illuminates a selected sample area (parallel electron beam) and detects the transmitted electron after passing through the samples	Provides detailed morphological properties (length, width, shape) of individual CNTs; high resolution can be used to distinguish between SWCNTs and MWCNTs; CNTs can be identified in energy filtered TEM images	Challenging sample preparation for tissues; it may be very hard to detect NPs in complex samples at low concentrations; low contrast (conventional TEM) due to reduced interactions between CNTs at the electron beam at high acceleration voltages
Thermal			0
CTO-375 ¹⁸	Quantification of carbon that remains after combustion at 375 °C for 24 h under excess air sample and subsequent chemical oxidation	Particularly good for complex matrices such as soil and sediment	Not fully tested for suspensions, requires high concentrations of CNTs and low concentrations of interferences (e.g., soot interfering with MWCNTs or graphene with SWCNTs)
Thermal Gravimetric Analysis (TGA) ^{20,171,172}	Quantification of mass percentage of phases with distinct thermal stabilities under a variety of reactive atmospheres (usually air) and relatively rapid temperature programs (e.g., heating rates of 5 C/min to 20 C/min,; room temperature- ca. 950 °C); each sample takes 1 h to 2 h total	A rapid technique that allows the quantification of multiple phases in a single sample, good for complex matrices, no special sample preparation needed	Effect of thermal ramp rate and reactive atmospheres on apparent phase distribution is not well understood (and is largely ignored), detection limits are relatively high for solid matrices, potential for interferences between sample matrix (e.g., other carbon nanomaterials, soot, or black carbon) and CNT decomposition temperatures
Thermal Gravimetric Analysis-Mass Spectrometry (TGA-MS) ¹⁴	TGA coupled with mass spectrometric detection of evolved gas fragments, typically in the 2 to 300 m/z range	Mass fragments can give insight into the chemical structure of the source material (e.g., C/H/O ratios or unique evolved fragments)	Current mass spectrometers have poor mass resolution (ca. 1 amu), relatively high detection limits, and low sampling rates relative to the chamber flush rate (i.e., consequently, only a small portion of the evolved mass is transferred to the MS); all reduce identification accuracy and increase detection limit

Total Organic Carbon (TOC) Analysis ⁷¹	TOC analysis can be conducted on water or soil samples by oxidizing (chemical, heated catalyst, UV) carbon to carbon monoxide or dioxide which is detected by infrared or other detectors	TOC analysis of waters has been used to measure CNTs in stock solutions in water	Very little optimization of temperature or catalytic conditions have been examined; its application to CNT stock solutions have been consistent with prepared masses; any organics, such as natural organic matter, in solution or soils would interfere; this is a non-specific method and thus matrices that contain sufficiently high concentrations of other carbon nanomaterials (e.g., graphene), soot, or black carbons would impact the technique
Thermal Optical Transmittance (TOT) ^{16,23}	As the sample is analyzed under programmed temperature, the volatilized and combusted carbon travels to an oxidizing oven, where it is transformed into carbon dioxide (CO ₂); the amount of elemental carbon is determined based on the CH ₄ signal measured using a flame ionization detector; sample is first heated under inert conditions to remove volatile organic carbon, then oxidizing carrier gas is used for elemental carbon; the portion of TC that is organic carbon or elemental carbon is defined by the method, which determines where the organic carbon-elemental carbon split is placed post-analysis; this split can be automatic on the basis of automatic optical correction; the optical transmittance or reflectance is observed throughout analysis, and the split is placed where the transmittance/ reflectance returns to the initial reading; for samples in which optical correction does not work, a manual split defined by the analyst should be used	Very reliable technique for detecting elemental carbon in environmental matrices, this technique could differentiate between types of CNTs based on their thermal stability	Too much organic carbon in a sample causes peak overlapping between elemental and organic carbon which affects the accuracy; similar carbonaceous materials such as graphene and fullerene will be counted in the CNT peak if they exist in the sample; unless the peak from CNT is far enough from other carbonaceous material, it is difficult to exclude the other carbonaceous materials but adjusting the temperature program might help to some extent
Isotopic labeling			
Carbon-13 Labelling ^{21,32,78}	A measure of the ratio of ¹³ C to ¹² C, applicable for all CNTs but works best for isotopically enriched or depleted CNTs	Instrumentation is readily available in many environmental laboratories	Highly dependent on matrix and large variability may be observed for CNTs that are not specifically ¹³ C enriched
Carbon-14 Labelling ^{15,26,30,99,124,152,173-181}	Measures beta emissions from carbon-14 emissions, can be used to quantify liquids after mixing with scintillation cocktail or any matrix after combustion in a biological oxidizer, autoradiography can provide spatial distribution of radioactivity	Provides definitive quantification of CNTs in complex matrices, can be used as an orthogonal technique to develop other analytical techniques, can be used to identify degradation products	High cost to synthesize radioactively labeled CNTs, safety concerns, limited availability of radioactively labeled CNTs

Other radioactive isotopes ⁹⁶⁻ 98	Measures release of emissions from a radioactive isotope that is associated (e.g., attached to a polymer wrapping the CNT) with the CNT	This approach can enable extremely low detection limits, can be used with a range of CNT surface functionalizations, non-destructive sample is possible for gamma emitters	Artifacts are possible if the radioactive isotope becomes separated from the CNT, it may be challenging or impossible to determine if this occurred in complex matrices without orthogonal CNT quantitation techniques
Additional Techniques			
Analytical Ultracentrifugation (AUC) 60,165,182-184	Measurement of sedimentation velocity distribution, can be used to determine particle density or size/shape distribution	Can measure entire CNT population via absorbance or interference measurement, high resolution, little size bias	Finicky technique that requires well understood and controlled samples for robust analysis
Gravimetric ¹⁸⁵	The CNT concentration in suspension is estimated by drying a fraction of the suspension and weighing it, or by determining the fraction of CNTs not suspended during the dispersion process (e.g., by sonication) by weighing the mass of CNT particles at the bottom of the container	Uses readily available equipment	Limited to high CNT concentrations, only applicable for aqueous suspensions
Microwave Method ^{62,125,186,187}	Measures the temperature rise of a sample at a specific microwave energy within a specific timeframe	Straightforward method for CNT detection and quantification in biological tissue, low cost	Not commercially available; it still remains to be investigated for environmental samples if interferences arise from other carbon allotropes with similar behavior in the microwave field (e.g., carbon black, soot) ¹⁸⁸
aF4-MALS ⁴⁶	Measures a shape factor (p=radius of gyration/hydrodynamic radius) of particles present in a complex liquid sample (e.g. surface water, leachate, soil and sediment extract), which is indicative of the particle aspect ratio; comparing these results to a CNT-free sample can then be used for CNT detection	Allows for CNT detection in water, soils, and sediments; may be useful in exposure studies	Need for the baseline of a CNT-free sample, full quantitative use currently not straightforward, often low CNT recoveries for aF4
Fable 7. Calestad to	abaiawaa faa CNT Quantitation		

$1342 \\ 1343 \\ 1344 \\ 1345 \\ 1346$ Table 2: Selected techniques for CNT Quantitation

Abbreviations: Asymmetric flow field flow fractionation with multi-angle light scattering (aF4-MALS), analytical ultracentrifugation (AUC), carbon nanotube (CNT), chemothermal oxidation at 375 °C (CTO-375), inductively coupled plasma-mass spectrometry (ICP-MS), near infrared fluorescence (NIRF), multiwall carbon nanotube (MWCNT), photoacoustic (PA), photothermal (PT), single particle inductively coupled plasma-mass spectrometry (spICP-MS), single-wall carbon nanotube (SWCNT), transmission electron microscopy (TEM), thermal gravimetric analysis (TGA), thermal gravimetric analysis-mass spectrometry (TGA-MS), total organic carbon (TOC), thermal optical transmittance (TOT).

N Carbon-14 Labelling Other radioactive isotopes Microwave method	UV/vis/near IR spectroscopy IR Fluorescence Spectroscopy/Microscopy Photoacoustic and Photothermal Raman spectroscopy spICP-MS Hyperspectral Imaging SEM, TEM, AFM TGA TOT C-13 Labelling AF4-MALS TGA-MS CTO-375	UV/vis spectroscopy Inorganic element analysis TOC Analysis Gravimetric Analysis
Custom built or requiring custom CNT synthesis	Commercially available but limited availability in environmental laboratories due to cost, expertise, etc.	Commercially available and broad availability in environmental laboratories
 Less available 		More available

Figure 1: Availability of CNT quantification techniques.

Abbreviations: Asymmetric flow field flow fractionation with multi-angle light scattering (AF4-MALS), analytical ultracentrifugation (AUC), chemothermal oxidation at 375 °C (CTO-375), near infrared fluorescence (NIR), single particle inductively coupled plasma-mass spectrometry (spICP-MS), thermal gravimetric analysis (TGA), thermal gravimetric analysis-mass spectrometry (TGA-MS), total organic carbon (TOC), thermal optical transmittance (TOT), scanning electron microscopy (SEM), transmission electron microscopy (TEM), atomic force microscopy (AFM).



Figure 2: Comparison between detection limits for analytical techniques in a water-only media under optimal conditions juxtaposed with a species sensitivity distribution for CNTs for acute toxicity testing of pelagic organisms. For the species

1357 sensitivity distribution, the 95 % confidence for the LC₅₀ values is shown by the gray shaded area around the curve. The

1358 detection limits for the techniques span a range of one order of magnitude (e.g., 1 mg/L to 10 mg/L). This figure is modified

1359 with permission from Garner et al.¹⁸⁹

- Abbreviations: Asymmetric flow field flow fractionation with multi-angle light scattering (aF4-MALS), analytical ultracentrifugation (AUC), chemothermal oxidation at 375 °C (CT0-375), near infrared fluorescence (NIRF), single particle inductively coupled plasma-mass spectrometry (spICP-MS), thermal gravimetric analysis (TGA), thermal gravimetric analysis-mass spectrometry (TGA-MS), total organic carbon (TOC), thermal optical transmittance (TOT).
- 1363



1366 Figure 3: Detection limits for analytical techniques in various media under optimal conditions and modeled environmental concentrations (1⁸⁴, 2⁸⁵, 3⁸⁶, 4⁷⁰, 5¹⁹); modeled environmental concentrations are not available for biological matrices. The 1367

1368 detection limits for individual techniques span a range of one order of magnitude (e.g., 1 mg/L to 10 mg/L).

1369 Abbreviations: Asymmetric flow field flow fractionation with multi-angle light scattering (aF4-MALS), analytical ultracentrifugation (AUC), chemothermal oxidation at 375 °C (CT0-375), near infrared

1370 1371 fluorescence (NIRF), single particle inductively coupled plasma-mass spectrometry (spICP-MS), thermal gravimetric analysis (TGA), thermal gravimetric analysis-mass spectrometry (TGA-MS), total organic carbon (TOC), thermal optical transmittance (TOT).

Supporting Information Quantification of carbon nanotubes in environmental matrices: Current capabilities, case studies, and future prospects

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Summary of contents

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					Release from polymer
Method Spectroscopic	Natural Waters	Cell Media	Biological Tissues	Soil/Sediment	nanocomposites
Absorbance ^{65,109,150,151}	Scattering	and absorbance of other components f	rom the matrix may interfer	e with CNT quantificatio	n
Near infrared fluorescence (NIRF) ^{15,29,31,152}	Scattering and absorbance of o	other components may interfere with Cl meas	NT quantification; backgrou urements	nd fluorescence of the m	natrix may also impact
Raman ^{13,122,123,153-164}	Humic acid, other Raman active organic contaminants, and suspended particles (e.g., clays) could impact the detection method as could background fluorescence	This matrix may have background fluorescence	This matrix may have auto-fluorescence and may limit light penetration	Light scattering by large particulate material, may require separation prior to Raman analysis	Presence of the aromatic compounds at high concentration could influence the signal as could fluorescence
Spectrometric				,	
Inorganic Elemental Analysis ^{29,32}	The background metallic content in the matrix should be clearly defined before ICP-MS analysis and compared to the metals present in the CNTs; the elemental specificity and the high matrix tolerance, makes ICP-MS based techniques practically independent to most common environmental and biological interferences				
Single particle inductively coupled plasma-mass spectrometry (spICP- MS) ²²	Difficult to distinguish CNTs from	other particulates containing the same	metals, this is most likely fo	r soils/sediments; extrac	tion may be needed first
Microscopic					
Atomic Force Microscopy ^{109,165}	For all matrices the presence of ar nanotube components; calculation or	ny other particulates depositing on the r f length distributions can be hindered b long nanotubes) and/or	neasurement substrate will y resolution issues (for short the presence of aggregates	require protocols for sel t nanotubes) and observ	ective removal of all non- ation bias (undercounting of
Hyperspectral Imaging ^{166,167}		For all matrices, soot and other black p	articles could impact the de	tection of CNTs	

Photoacoustic (PA) and photothermal (PT) ^{24, 168-} ¹⁷⁰	Water is a good PA/PT solvent, anythir scatters the beam(s) would decrease s	ig else in the sample that absorbs or ignal or increase background	PA/PT works well in tissues transparent to beam(s); PT sensitivity drops in non- transparent tissue	No reports in literature; would be a difficult matrix to detect CNT with a lot of scattering and absorption of the beam(s)	A polymer matrix does not inhibit CNT detection; as long as there is a thermal response in the matrix, PT/PA can detect a signal
Scanning Electron Microscopy, Scanning Transmission Electron Microscopy, Transmission Electron Microscopy (TEM) ^{27,66}	Biopolymers, low concentration of CN to overlapping particles on the sample	Γs compared to other particles leading s holder	Other fibrillar particles; low contrast between CNTs and biological tissue	Other fibrillar particles; compared to other par particles on the sample	; low concentration of CNTs ticles leading to overlapping es holder
Thermal CTO-375 ¹⁸	Very little interference in this matrix except for high N, organic carbon, or black carbon content waters	Matrices with high N or organic C con higher stability materials that, togeth concentrations, can interfere with the conversely some matrices can produc reduce the oxidation temperature of	itent can char and form er with high BC e analysis of CNTs; se "catalytic" effects that recalcitrant carbons	Sample specific oxidative strength (protective or catalytic) leading to variable recoveries of spiked CNTs; high organic C can char and high BC content can interfere with the CNT analysis	Some tested polymers (e.g., gamma-poly caprolactone) have lower thermal stability than CNTs which makes this a promising approach
Thermal Gravimetric Analysis (TGA) ^{20,171,172}	Very little interference expected except for waters with levels of BC that are approximately equal in concentration to the CNTs, NOM can stabilize CNTs	Isolated test materials show little inte testing needed, peptone also binds to oxidation temperature	erference, full matrix o CNTs and may change	Isolated test materials show little interference; full matrix testing needed, major	Interferences are unclear, tested epoxies have overlapping thermal stabilities with CNTs, and seem to influence the

				challenge in sample size (typically > 10 mg) and overlap of oxidation temperatures may hinder detection of CNTs	burn temperature of one another; overlap of oxidation temperatures or changes in thermal stability of CNTs in this matrix can hinder CNT quantification; in some cases, CNTs can be extracted from the matrix to give clear ratios of matrix:CNT composition
Thermal Gravimetric Analysis-Mass Spectrometry (TGA-MS) ¹⁴	Very little interference; interference with BC can be reduced by MS peak deconvolution	Interferences with this matrix are unknown, but the unique chemical composition of cell media is promising for lower interferences	Isolated test materials show little interference and the unique chemical composition of biological tissues suggests low interferences, major challenge in sample size and overlap of oxidation temperatures may hinder detection of CNTs	Few direct interferences, but can raise background levels and raise the detection limit	The unique chemical compositions of most polymers suggests low interferences, evolved gases should be distinct for CNTs as compared to the polymer matrix
Total Organic Carbon (TOC) Analysis ⁷¹	Interference exists from any organ	ic matter (natural organic matter, soil o polym	organic matter, cellular mate ners, etc.)	erial, serum or other orga	anic compounds, organic

Thermal Optical Transmittance (TOT) 16,23

Typically little interference

Interferences should be minimal, but may arise if cell material chars into optically absorptive or thermally stable material

Interferences can be minimized by preparatory digestions (demonstrated for mouse lung)

Few direct interferences, unless the soil or sediment has a high non-CNT organic load

Potentially interfering, as many polymers will exhibit poor degradability under the inert atmosphere utilized in the first phase of this method; This could cause charring and confound the measurement of EC once the oxidative atmosphere is introduced

Isotopic labeling

Carbon-13 Labelling ^{21,32,78}	Very little interference expected	Very little interference expected	Separation of CNTs from tissues is advised as accumulation of CNTs may be tissue dependent; background δ^{13} C signatures are necessary for each tissue type	Sulfates may interfere with the preparation of pure CO_2 ; CNT-free background required for comparative $\delta^{13}C$ signatures	CNT-free background require for comparative δ^{13} C signatures
Carbon-14 Labelling ^{15,26,30,99,124,152,173} -181	The potential biases depend on how th compounds may interfere with scintilla emissions and could lead to autofluore expected if the sample is combusted u	ne carbon-14 is quantified; some ation cocktails adsorbing beta escence; these issues would not be sing biological oxidation	It may be possible to sonicate the tissue in liquid scintillation fluid, but there may be incomplete dispersion of the CNTs or quenching of the radioactivity; interferences are	Interferences have not been observed in previous studies with biological oxidation of the samples; good recovery was also found when sonicating SWCNTs with sodium dodecyl sulfate and using	There may be interference from quenching if the sample is added to liquid scintillation cocktail, but interferences would not be expected for biological oxidation

unlikely with biological	liquid scintillation
oxidation	counting

Other radioactive isotopes ⁹⁶⁻⁹⁸	This would depend to some extent on the radioactive isotope added and quantificatic different matrices; however, the stability of the radioactive tracer may be impacted tissues	s would depend to some extent on the radioactive isotope added and quantification used but generally interferences would not be expected for these Jifferent matrices; however, the stability of the radioactive tracer may be impacted by the dispersion process in the matrix or metabolic processes in tissues			
Additional					
Techniques					
Analytical	Measurement of sedimentation requires homogenous dispersions with measureable	viscosities and densities; significant light scattering from suspended			
Ultracentrifugation (AUC) ^{60,165,182-184}	particles from the matrix will additional likely complicate all b	particles from the matrix will additional likely complicate all but the most rigorous experimental protocols			
Gravimetric ¹⁸⁵	Measurement of CNTs in these matrices would encounter significant biases depend on the mass of other compounds that would be deposited when	This technique is not applicable for these matrices			
	drying samples except at very high CNT concentrations				
Microwave Method ^{62,125,186,187}	Other carbon forms such as soot may cause interferences; this interfe	Other carbon forms such as soot may cause interferences; this interference would be most likely for soils and sediments			
aF4-MALS ⁴⁶	No known interferences, but theoretically other low density fibre-like/high aspect ratio particles may interfere; if these particles exhibited lower thermal stability compared to the CNTs, oxidation could potentially be used to selectively remove them				

Table S1: Potential Matrix Interferences for Selected Techniques for CNT Quantitation

Abbreviations: Asymmetric flow field flow fractionation with multi-angle light scattering (aF4-MALS), analytical ultracentrifugation (AUC), black carbon (BC), carbon nanotube (CNT), chemothermal oxidation at 375 °C (CTO-375), inductively coupled plasma-mass spectrometry (ICP-MS), near infrared fluorescence (NIRF), photoacoustic (PA), photothermal (PT), single particle inductively coupled plasma-mass spectrometry (spICP-MS), single-wall carbon nanotube (SWCNT), thermal gravimetric analysis (TGA), thermal gravimetric analysis-mass spectrometry (TGA-MS), total organic carbon (TOC), thermal optical transmittance (TOT).

Method	Impact of CNT agglomeration	Impact of CNT Oxidation/Degradation	Wrapping with Organic Molecules (proteins, NOM)
<i>Spectroscopic</i> Absorbance ^{65,109,150,151}	Measured absorbance signal per unit mass typically decreases above a threshold level of nanotube aggregation, especially for the intrinsic nanotube optical transitions, although apparent absorbance in the UV and visible regions may broadly increase due to increased light scattering by larger particles; for intrinsic optical transitions, the transition wavelength will typically red shift and peak intensities will decrease with any reduction from individualized dispersion	Absorbance of intrinsic optical transitions typically decreases monotonically above very low levels	In the absence of changes in agglomeration state, the adsorption of material to the nanotube interface generally will affect the absorbance mostly through red/blue shifts in intrinsic optical transition wavelengths by modification of the local dielectric environment; changes to the surface accessibility of the bulk solvent can also affect optical transition intensities
Near infrared fluorescence (NIRF) 15,29,31,152	Peak shifts and intensity decrease for SWCNTs could oc	cur for either of these changes	Variable
Raman ^{13,122,123,153-164}			
Absorbance ^{65,109,150,151}	Not a significant factor but G and D band ratio may be sensitive to sample agglomeration	Raman spectra are very sensitive to oxidation or degradation	Vibrational features are sensitive to structural stress which may be caused by wrapping with organic molecules or polymers
Spectrometric			. ,
Inorganic Elemental Analysis ^{29,32}	Minimum impact on elemental analysis when a complete sample digestion is performed	Any loss of metals intercalated in CNTs before the elemental analysis would lead to biased results	Minimum impact when the wrapping does not alter the elemental composition of CNT
Single particle inductively coupled plasma-mass spectrometry (spICP- MS) ²²	Severe undercounting effect on actual CNT concentrations since each agglomerate may only be counted as one single pulse depending on the dwell time	Important influence on sizing and counting results because of the increasing contribution of smaller CNTs containing metal masses below instrument detection limit	Wrapping would affect physical transport of the CNT in introduction system, increasing the uncertainty on the size and number concentration determination
Microscopic			
Atomic Force Microscopy ^{109,165}	Agglomeration or oxidation/degradation may impact appa	arent size distribution and hinder	Variable

analysis

Hyperspectral Imaging ^{166,167}	Better optical visibility due to enhanced scattering from	Potential changes in spectral pro	ofiles from oxidation/degradation or
	agglomerates	wrapping with	organic molecules
Photoacoustic (PA) and Photothermal (PT) ^{24, 168-170}	Will follow the same changes that affect Absorbance; anyt	hing that changes the absorption of	of the CNTs would affect the PT/PA
	signal causing shifts in peak wavelength and changes in ab	sorption cross-section; degradation	n would certainly affect the PT/PA
	signal; the effect of agglomeration, oxidation, and the add	ition of physisorbed or chemisorbe	ed ligands would be case-by-case
Scanning Electron Microscopy, Scanning Transmission Electron Microscopy, and Transmission Electron Microscopy ^{27,66}	CNTs will still be detected if investigated manually, but automated analysis may fail to identify CNT in agglomerates	Change in the size distribution, depending on the extent of degradation	May reduce the image resolution due to contamination effects (resulting from the volatilization and redeposition of the organic

Thermal CTO-375¹⁸

CNT agglomeration will slightly increase thermal stability, but not to an extent discernable by CTO-375

Oxidation and degradation reduce CNT thermal stability, which would enhance separation from BC but require a different cut off temperature to quantify SWCNTs; MWCNTs will still be interfering with BC Organic coatings should be resolved (e.g., more labile than CNTs) by CTO-375 and not affect the measurement; however, proteins can char and cause interference

material under the electron

beam)

Thermal Gravimetric Analysis (TGA) 20,171,172	CNT thermal stability will increase measurably; SWCNTs may no longer be resolved from BC or soot; MWCNTs should still have higher thermal stability than BC; higher temperature shoulders on oxidation peaks occurs with bundling, changes in oxidation temperature of material when bundled vs not	Oxidation and degradation of CNTs will reduce the thermal stability, which should help resolve SWCNTs from soot, but will likely not change the MWCNT thermal stability to such an extent that it interferes with BC	Organic coatings can influence the thermal stability of the CNTs, where lower onset temperatures and broader mass loss events have been observed, increasing potential interferences; proteins can char and cause interference
Thermal Gravimetric Analysis-Mass Spectrometry (TGA-MS) ¹⁴	CNT thermal stability will increase measurably, no anticipated change in MS signal	Oxidation of CNTs will reduce the MS-derived advantages, which leverage the low-oxygen content of CNTs; potential for changes in decomposition products	Organic coatings can change CNT thermal stability and should increase the O and/or N content of the diluting matrix; thus, CNT- derived depletions in O would become easier to observe with organic matter coatings
Total Organic Carbon (TOC) Analysis ⁷¹	Unlikely to be impacted by aggregation	Oxidation of CNTs would likely improve detection, as TOC analysis relies upon complete conversion to gaseous carbon mon- or di-oxides	Any organic surface coating (citrate, amine, etc.) contributes to the carbon detected from the CNT
Thermal Optical Transmittance (TOT) ^{16,23}	Sample on the filter won't be uniform, the split point of organic carbon/elemental carbon needs to be manually chosen instead of by optical information	Oxidation/degradation decreases the thermal stability and causes peak position shift; no issue with quantifying CNTs unless the sample has huge amount of organic carbon and the peak position of CNT after shifting is getting close to organic carbon	Having too much organic carbon may affect the thermal stability of CNT, and the signal from organic carbon will overlap that of elemental carbon; organic carbon should be resolved as much as possible
Isotopic labeling			
Carbon-13 Labelling ^{21,32,78}	Comprehensive oxidation of the CNTs required to prevent isotopic fractionation; agglomeration may affect thermal stability in closed-tube-combustion approaches, and efforts should be made to ensure complete combustion	Pre-analysis CNT oxidation may have slight impacts on CNT δ^{13} C signature (by virtue of reactive fractionation); these should be small depending on the extent of surface oxidation and/or if that process removes CNT-C from the CNT matrix	Wrapping with organic molecules will affect δ 13C; the effect will depend on the δ 13C of the molecule, and measures to separate the coating from the CNT are critical

Carbon-14 Labelling ^{15,26,30,99,124,152,173-181}	The impact of agglomeration would depend on the quantification procedure used; interference from self- quenching has been reported in some studies with agglomerates of CNTs, but this would not be expected for quantification using biological oxidation	This is not expected to impact this approach; carbon-14 analyses of released carbon dioxide has been used to quantify CNT degradation	This may impact measurements with liquid scintillation counting of dispersed CNTs but should not impact samples combusted using biological oxidation
Other radioactive isotopes ⁹⁶⁻⁹⁸	This would not be expected to impact most isotopes unless self-quenching occurs	Oxidation or degradation may render this technique unusable if these processes lead to substantial separation of the radioactive isotope from the CNT	This would not be expected to impact most isotopes unless quenching occurs
Additional Techniques			
Analytical Ultracentrifugation (AUC) 60,165,182-184	Aggregates rather than primary particle would be measured, data analysis potential decreased	Interpretation of results will become suspect due to differences in actual sample with respect to expected behavior	Variable effects, will likely bias size analysis
Gravimetric ¹⁸⁵	No impact	No impact unless there is complete degradation to CO ₂	This can limit the accuracy of this approach since the concentration of these organic molecules will need to be assumed to be homogeneously distributed
Microwave Method ^{62,125,186,187}	These potential interference	s have not been tested for this tech	mique
aF4-MALS ⁴⁶	Enhances material losses on the membrane and hinders accurate shape factor determination; thus agglomeration has to be avoided	Modifies CNT interactions with higher or lower losses depending ma	the membrane which can lead to g on carrier solution and membrane iterial

Table S2: Potential Interferences for CNT Quantitation from Changes to CNTs for Selected Techniques

Abbreviations: Asymmetric flow field flow fractionation with multi-angle light scattering (aF4-MALS), analytical ultracentrifugation (AUC), black carbon (BC), carbon nanotube (CNT), chemothermal oxidation at 375 °C (CTO-375), multiwall carbon nanotube (MWCNT), near infrared fluorescence (NIRF), photoacoustic (PA), photothermal (PT), single particle inductively coupled plasma-mass spectrometry (spICP-MS), single-wall carbon nanotube (SWCNT), thermal gravimetric analysis (TGA), thermal gravimetric analysis-mass spectrometry (TGA-MS), total organic carbon (TOC), thermal optical transmittance (TOT).

Method	Relevant reference materials or standard methods	Key steps in instrument calibration	Challenges to standardization (i.e., traceability to the SI)
Spectroscopic			
Absorbance	Informational values RM8281, ISO/TS 10868:2011	Calibration of wavelength and intensity performed at 100 % transmittance. Usable wavelength range should be established by testing the absorbance of a blank sample and considering regions of high absorbance, scattering interference and Beer's law considerations	Chemical environment can affect intensity and peak positions. Absorbance of water in NIR wavelengths for cell path lengths > 1mm.
Near infrared fluorescence	Informational values RM8281, ISO/TS 10867:2010	Traceable lamp detector train calibration	Chemical environment can affect intensity and peak positions in-filter effects
Photoacoustic and photothermal	CNTs with well characterized absorbance of narrow size distribution in a pure solvent could be used for calibration	A standard sample with material similar to the sample CNTs (that absorbs the same wavelength) can be used to tune the setup. Laser power, sensitivity, and time constant can be adjusted for the sample as needed	No standards published or referenced to date. Short shelf life for samples if in situ CNTs degrade or change over time. Difficult traceability to SI.
Raman	Frequency (x-axis) calibration standards ASTM E1840, Intensity (Y- axis) E2911, E2529, NIST SRM series 224X	Choose the appropriate standard for frequency and intensity depending on the excitation wavelength. Alternatively, a series of standard solutions (dilution series) of the pure analyte in combination with the internal standard can be used	
Spectrometric		standard can be used	
Inorganic Element Analysis	Dissolved standards of the monitored elements are required to determine the instrument sensitivity for the elemental quantification. Potential influences from residual carbon content and dissolved solids can be accounted for by suitable calibration techniques, including isotope dilution, matrix matched standards and the method of additions. SRM 2483 (single-wall carbon nanotubes (raw soot)) could be used to test instrument performance	ISO/TS 13278:2011E, This Technical Specification provides reference standard methods for the determination of elemental impurities in CNTs using ICP-MS. Results traceable to the SI can be readily achieved using traceable high- purity calibration standards. Calibration is performed with solutions having known concentrations of the metallic analytes of interest and matrix- matched to the composition of the prepared samples.	Lack of control environmental and biological matrices. Guarantee that sample digestion is quantitative prior to elemental analysis.

Single particle inductively coupled plasma-mass spectrometry (spICP- MS)	Reference Materials: Single element standard solutions available from numerous reference materials producers; NIST RM 8013 Gold Nanoparticle, Nominal 60 nm Diameter; Standards: ISO TS13278 Determination of metal impurities in samples of carbon nanotubes using inductively coupled plasma mass spectrometry	Calibration of ICP-MS instrument sensitivity is performed with solutions having known concentrations of the metallic analytes of interest and matrix-matched to the composition of the prepared samples. Calibration of sample transport efficiency is performed using metallic nanoparticles having known size (metal does not need to be the same as the trace metal analytes). Sample transport efficiency may also be calibrated using the waste collection method, but this method is generally less reliable.	Reference CNT samples with homogeneous size and controlled metal impurities contents are required to address the standardization.
Microscopic			
Atomic Force Microscopy	ASTM E2859-11 Standard guide for size measurement of nanoparticles Using atomic force microscopy, NIST RM8281, NRC Canada SWCNT-1	In-plane resolution, i.e. distance/pixel should be selected to enable identification of smallest expected particles of interest.	Surface roughness of deposition substrates varies significantly with preparation methodology. Polydisperse samples may require measurements at multiple resolutions to identify small particles, and to locate larger particles.
Hyperspectral Imaging		A representative spectral library is generated from the parent material. The spectral library is then used to detect the same material in a sample (e.g. cell) using a mapping algorithm	Unspecific absorption in the VNIR spectral range
Scanning Electron Microscopy	ISO/TS 10798:2011 Nanotechnologies Charaterization of single-wall carbon nanotubes using scanning electron microscopy and energy dispersive X-ray spectrometry analysis	Reference materials (regarding CNTs number concentrations) must be use to evaluate instrumental losses during sample preparation	Reference CNT suspensions with certified number concentrations must be developed; the shelf life of these suspensions maybe limited due to CNT agglomeration
Transmission Electron Microscopy	ISO/TS 10797:2012 Nanotechnologies Characterization of single-wall carbon nanotubes using transmission electron microscopy	Reference materials (regarding CNTs number concentrations) must be use to evaluate instrumental losses during sample preparation	Reference CNT suspensions with certified number concentrations mus be developed. The shelf life of these suspensions maybe limited due to CNT aggregation
Thermal			
СТО-375			
Thermal Gravimetric Analysis	ISO/TS 11308:2011 Nanotechnologies Characterization of single-wall carbon nanotubes using thermogravimetric analysis	Temperature and mass calibration required	No reference materials for temperature calibration; traceable mass standards available
Thermal Gravimetric Analysis-Mass Spectrometry		Temperature and mass calibration required. MS peak identification database needed.	No reference materials for temperature calibration; traceable mass standards available
Total Organic Carbon Analysis			
Thermal Optical Transmittance	NIOSH, Elemental Carbon (Diesel Particulate): Method 5040. In NIOSH, Manual of Analytical Methods, 4th ed.; 2003.	Sucrose solution and methane gas carbon standandards are often used for mass calibration	Each CNT has slightly different peak position depending on defect, purity, functional group etc.; also it differs by the temperature program

Isotopic labeling			
Carbon-13 Labelling	Standards include calcium carbonate (commonly used Vienna PeeDee Belemnite and NBS ¹³ C standard), barium carbonate	Ensuring accuracy of standards is key to reliable measurements. Also, running standards throughout sample analysis is required to understand measurement drift.	Samples containing sulfate cause contamination in the final product. Small samples may not release enough gas for the analysis. Nanotubes will differ in their 13C ratios based on original source of C
Carbon-14 Labelling	Standards include NIST 4222C (carbon-14 hexadecane radioactivity standard solution)	Calibration depends on the method (liquid scintillation counting, autoradiography, biological oxidation) used to quantify the radioactivity. For all methods, it may be necessary to calibrate the instrument using other radioactive carbon-14 chemicals or elements.	
Other isotopes	Multiple radioactivity standards are available from NIST (e.g., 4915F cobalt-60 radioactivity standard solution) and from other organizations	Calibration depends on the instrument used to measure the radioactivity	One principal challenge is the stability of the radioactive isotope onto the carbon nanotube.
Additional techniques	-		
Analytical Ultracentrifugation	SRM under development for ensuring radial measurement precision; sedimentation of Bovine Serum Albumin (BSA) frequently	External evaluation of temperature calibration and bulk solution viscosity and density properties are critical	Requires unique absorbance or refractive index signals from solute differentiable from media.
Gravimetric	A broad range of mass RMs and protocols are available for gravimetric measurements	Balances can be calibrated using device-specific procedures, reference masses are readily available	Works only for a limited number of conditions and matrices
Microwave Method	CNT material used in the exposure experiment, (reference) control material (e.g., CNT-free biological tissue such as NIST SRM 1573 Tomato leaves)	The very same CNT material that is to be quantified must be used to calibrate the instrument; a calibration curve is generated using the thermal response as a function of known CNT amounts spiked into tissue samples	A main limitation to standardization is that the instrument used to make these measurements is not readily available
Field flow fraction/asymmetric flow field flow fraction/asymmetric flow field flow fraction- multi-angle light scattering	Certified polystyrene (PS) beads (Single or mix, available from NIST or other sources). Bovine serum albumin (BSA). Any other certified particle standard (e.g., Au, Ag, and SiO2) that can be dispersed in the carrier solution	PS beads dispersed in the used carrier solution are used for retention time calibration (hydrodynamic diameter). An isotropic scatterer is used for normalization of the MALS detector angles (e.g. 20nm PS beads). BSA is used for molecular weight calibration of the MALS detector	Reference CNT samples with homogeneous size and controlled particle impurities (e.g. soot) would be required for aF4-MALS quality assurance. Changes in the chemical environment of the CNTs as well as changes of the CNTs themselves (e.g., surface functionalizations, length distributions) can affect retention time in aF4; standardized methods must

Table S3: Standards and calibration of selected carbon nanotube quantification techniques

include extensive methodological details to ensure

reproducibility

Material	NIST SRM 2483 Single- wall carbon nanotube soot	NIST RM8281 Single-wall carbon nanotubes (dispersed, three length- resolved populations)	NRC Canada: SWCNT-1 Single- wall carbon nanotube certified reference material	JRC Multiwall carbon nanotube representati ve test materials
Reference Material	Yes	Yes	Yes	No

(Y/N)

Are certified, reference, or information values provided for these characteristics?

AFM imaging	No	Information values (length distributions)	Information values	No
Elemental composition	Certified, reference, and information values	No	Certified and reference values	Yes
NIR fluorescence spectra	No	Information values	Information values	No
Raman ratio	Reference values	Reference values	Reference values	Yes
Raman spectra	Information values	Information values	Reference values	Yes
SEM imaging	Information values	No	Information values	No
Specific surface area	No	No	Reference values	Yes
TEM imaging	Information values	Information values	Information values	Yes
Thermogravimetric analysis (residual mass and oxidation temperature)	Reference values	No	Reference values	Yes
UV-vis-NIR absorbance spectra	Information values	Information values	Information values	No
X-ray diffraction	No	No	Reference values	Yes

Table S4: Characterization of carbon nanotube reference materials, standard reference materials, and representative test material.